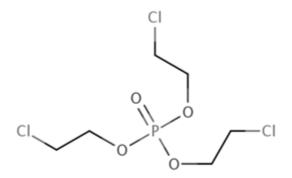
## **Risk Evaluation for** Tris(2-chloroethyl) Phosphate (TCEP)

## **Supplemental Information File:**

### **Benchmark Dose Modeling Results** for TCEP

**CASRN: 115-96-8** 



September 2024

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# ABBREVIATIONS AND ACRONYMS

280	AIC	Akaike information criterion
281	BMD	Benchmark dose
282	BMDL	Benchmark dose lower bound
283	BMR	Benchmark responses
284	CSF	Cancer slope factor
285	EPA	Environmental Protection Agency
286	ER	Extra risk
287	ICR mice	Institute for Cancer Research mice
288	LOAEL	Lowest-observed-adverse-effect-level
289	NOAEL	No-observed-adverse-effect-level
290	POD	Point of departure
291	RD	Relative deviations
292	SD	Standard deviation

#### 1 BENCHMARK DOSE MODELING RESULTS

The U.S. Environment Protection Agency (EPA or the Agency) performed benchmark dose (BMD) modeling using EPA's BMD modeling software (BMDS), Version 3.2.0.1) for the health domains that were identified during hazard identification and that received a judgment of likely ("evidence indicates that TCEP exposure likely causes [health effect]") and suggests ("evidence suggests but is not sufficient to conclude that TCEP exposure causes [health effect]") during evidence integration. EPA considered that TCEP is likely to cause the following health endpoints for which BMD modeling is presented: neurotoxicity, reproductive toxicity, kidney toxicity, and cancer (kidney tumors). EPA considered that TCEP exposure results in a suggests conclusion for: mortality, liver toxicity, and developmental toxicity. EPA conducted BMD modeling in a manner consistent with EPA's *Benchmark Dose (BMD) Technical Guidance* (U.S. EPA, 2012).

EPA used dichotomous models to fit quantal data (*e.g.*, incidences of karyomegaly) and continuous models to fit continuous data (*e.g.*, kidney weights), as recommended by EPA's *BMD Technical Guidance* (<u>U.S. EPA, 2012</u>). The BMD/BMDLs are provided based on a daily exposure (*i.e.*, seven days per week) for easier comparison across all hazard endpoints and thus, doses were adjusted as needed before BMD modeling. EPA modeled endpoints that had statistically significant pairwise comparisons between individual doses and controls or significant dose-response trends. EPA also considered potential biologically significant changes from controls where possible and/or changes that appeared to exhibit a dose-response relationship upon visual inspection. Multiple health endpoints may have been modeled from each study, depending on the relevance of the data to adverse health outcomes and to identify sensitive health endpoints for each domain.

Although some of the data sets could be fit using models after dropping doses (either 1, 2, or 3 of the highest doses), EPA considered only modeling results from full data sets for use in quantifying risk. This document does not present results of modeling exercises in which none of the models in the BMD suite provided an adequate fit to the full data sets. Several additional endpoints evaluated in various TCEP toxicity studies were not considered for BMD modeling because the changes were observed only at the highest dose. Studies were also not considered for BMD modeling if the lowest-observed-adverse-effect-levels (LOAELs) were more than 10-times greater than the most sensitive LOAEL for the health domain. If BMD modeling was not possible or when data did not fit the available models, EPA used no-observed-adverse-effect-levels (NOAELs) and LOAELs during point of departure (POD) selection for the risk evaluation.

EPA relied on the BMD guidance and other information to choose benchmark responses (BMRs) appropriate for each endpoint. Although the *BMD Technical Guidance* doesn't recommend default BMRs, it describes how various BMD modeling results compare with NOAEL values, and the guidance recommends calculating 10 percent extra risk (ER) for quantal data and one standard deviation (SD) for continuous data to compare modeling results across endpoints. EPA also modeled percent relative deviations (RD) for certain continuous endpoints. EPA's choice of BMRs for the TCEP health endpoints is described in more detail in the following sections that present BMD modeling results for each health domain.

When modeling dose-response relationships, the data can be modeled as either ER or additional risk. EPA modeled the data as ER. EPA's *BMD Technical Guidance* defines extra risk (ER) as "a measure of the proportional increase in risk of an adverse effect adjusted for the background incidence of the same

effect." Mathematically, extra risk is equal to [P(d) - P(0)]/[1 - P(0)]. P(d) is the probability of the effect at dose d, and P(0) is the probability of risk with no exposure to a hazard (U.S. EPA, 2012).

Of the modeled BMDLs, critical endpoints and their PODs used as the basis of risk estimates are decreased numbers of seminiferous tubules (Section 1.1.2.1), changes in path length in the Morris water maze (Section 1.1.1.1) and increased incidence of renal tubule adenomas and carcinomas (Section 1.2.1).

#### 1.1 Non-cancer

#### 1.1.1 Neurotoxicity

#### 1.1.1.1 Path Length in the Morris Water Maze Test in Female Rats

Path length in the Morris water maze test decreased in female rats exposed to TCEP for 60 days (Yang et al., 2018). First, the administered doses were duration adjusted to estimate an equivalent oral dose for animals exposed for seven rather than five days per week. Then, dichotomous models were used to fit dose response data.

A BMR of 1 SD, 10, 20, and 30 percent relative deviations were modeled according to EPA's *Benchmark Dose Technical Guidance* (<u>U.S. EPA, 2012</u>). EPA chose the BMR of 20 percent RD as the most appropriate measure of relevant biological change (<u>U.S. EPA, 2022</u>) when comparing with other PODs. The doses and response data used for the modeling are presented in Table 1-1.

Table 1-1. Path Length Decreased in the Morris Water Maze Test Selected for Dose-Response Modeling for TCEP from a 60-Day Study

Dose (mg/kg/day)	Number of Animals	Mean	SD
0	10	685	144.90
50	10	602	106.12
100	10	470	114.28
250	10	317	110.20

The BMD modeling results for path length in the Morris water maze test are summarized below in Table 1-2. The constant variance model provided an adequate fit to the variance data. With the constant variance model applied, all models except for the Exponential 5 and Hill models provided adequate fit to the means. The BMDLs for the fit models were sufficiently close (differed by <3-fold); therefore, the model with the lowest AIC was selected. The Exponential 2 and 3 models converged on the same model and had the lowest AIC; the Exponential 2 model is the more parsimonious choice.

<sup>&</sup>lt;sup>1</sup> EPA's *BMD Technical Guidance* also uses the terms excess incidence and excess risk, which are defined more generally as increased risk or incidence above control or background responses. These terms can refer to either additional or extra risk (U.S. EPA, 2012).

Table 1-2. Summary of BMD Modeling Results for Path Length in the Morris Water Maze Test in Female Rats in the 60-Day Study

	Goodness of (Means	(Mean		BMD 1SD	BMDL 1SD	BMD 10%RD	BMDL 10%RD	BMD 20%RD	BMDL 20%RD	BMD 30%RD	BMDL 30%RD	Basis for Model
Model	Test 4 p-value	AIC	(mg/kg- day)	Selection								
<b>Exponential 2</b>	0.636723	499	57	41	33	26	69	55	111	87		
Exponential 3	0.636723	499	57	41	33	26	69	55	111	87		
Exponential 4	0.394512	501	50	29	29	18	62	40	102	68	The Exponential 2 and Exponential 3 converged on the same model and had the lowest AIC; the Exponential 2 model is the parsimonious choice.	
Exponential 5	N/A	502	61	32	44	19	70	42	96	71		
Hill	N/A	502	61	31	45	18	69	41	96	70		
Polynomial Degree 3	0.298849	501	80	62	46	39	91	77	137	116		
Polynomial Degree 2	0.298849	501	80	62	46	39	91	77	137	116		
Power	0.298849	501	80	62	46	39	91	77	137	116		
Linear	0.298849	501	80	62	46	39	91	77	137	116		

Plots of the Exponential 2 model with BMRs of one SD, or 10, 20, or 30 percent RD are shown in Figure 1-1, Figure 1-2, Figure 1-3 and Figure 1-4, respectively. Additional modeling details, including model parameters, goodness of fit at each dose, and log likelihood are shown below in Figure 1-5.

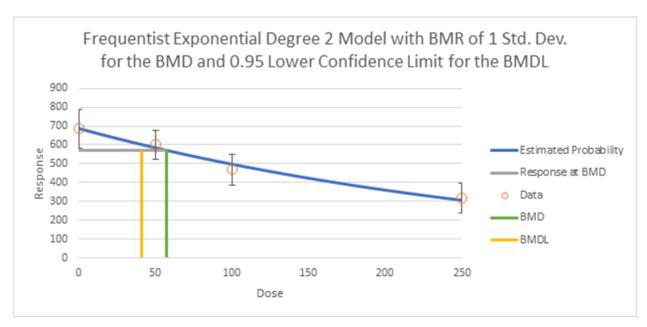


Figure 1-1. Plot of Response by Dose with Fitted Curve for the Selected Model (Exponential 2) for Path Length in the Morris Water Maze Test in Female Rats Exposed to TCEP Via Oral Gavage (60-Day Study) and BMR of 1SD (Constant Variance Assumed)

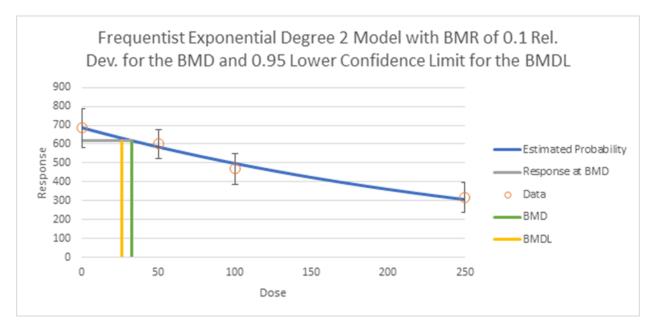


Figure 1-2. Plot of Response by Dose with Fitted Curve for the Selected Model (Exponential 2) for Path Length in the Morris Water Maze Test in Female Rats Exposed to TCEP Via Oral Gavage (60-Day Study) and BMR of 10 Percent Relative Deviation (Constant Variance Assumed)

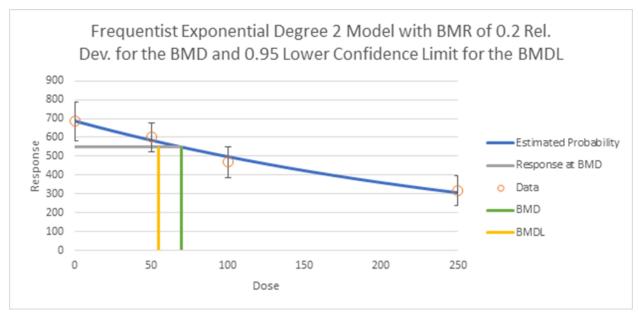


Figure 1-3. Plot of Response by Dose with Fitted Curve for the Selected Model (Exponential 2) for Path Length in the Morris Water Maze Test in Female Rats Exposed to TCEP Via Oral Gavage (60-Day Study) and BMR of 20 Percent Relative Deviation (Constant Variance Assumed)

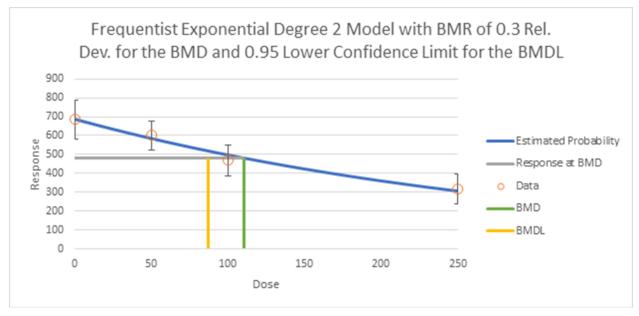


Figure 1-4. Plot of Response by Dose with Fitted Curve for the Selected Model (Exponential 2) for Path Length in the Morris Water Maze Test in Female Rats Exposed to TCEP Via Oral Gavage (60-Day Study) and BMR of 30 Percent Relative Deviation (Constant Variance Assumed)

Be	nchm	ark Dose	1						
BMD	_	.99372292	1						
BMDL	-	.99542231	1						
BMDU	_	7.274695	1						
AIC	_	99.106493	1						
Test 4	_		1						
P-valu	e 0.6	636722823							
D.O.F.		2	1						
			_						
- 1	Model	Parameters		1					
# of				1					
Param	eters		3						
Var	iable	Estima	te	]					
	а	685.5495	5504	]					
	b	0.003221	1614						
Good	Iness								
of									
n	01	Estimated	i	Calc'd	Observed	Series and SD	Calc'd	Observed	C1- 4 D14
Dose	Size	Median		Median	Mean	Estimated SD	SD	SD	Scaled Residu
0	10	685.549550	04 (	684.94	684.94	114.994889	144.9	144.9	-0.01676220
50	10	583.555800	03 (	501.76	601.76	114.994889	106.12	106.12	0.50060254
100	10	496.736336	53 4	469.62	469.62	114.994889	114.28	114.28	-0.74567996
250	10	306.377274	19	317.03	317.03	114.994889	110.2	110.2	0.29294236
Likeli	hoods	of Interest				_			
				# <u>of</u>					
	_	Likelihood*	Para	meters	AIC				
A1	_	16.1018257		5	502.2036	_			
A2	_	15.4659596		8	506.9319				
A3	-	6.1018257	_	5	502.2036	_			
fitted	_	16.5532465		3	499.1064				
R	-26	54.4339647		2	532.8679	929			
Te	ests of	fInterest							
	-2*Lc	g(Likelihood	Tes	t					
		Ratio)	df	p-va	alue				
Test		93601035	6	<0.0	0001				
Test	37.								
$\overline{}$		71732238	3	0.735	85623				
$\overline{}$	1.2	71732238 71732238	3	_	85623 85623				

Figure 1-5. Details Regarding the Selected Model (Exponential 2) for Path Length in the Morris Water Maze Test in Female Rats Following Oral Exposure to TCEP in a 60-Day Toxicity Study

#### 1.1.1.2 Necrosis of the Neurons of the Hippocampus in Female Rats

Increased necrosis of the neurons of hippocampus was observed in female rats exposed to TCEP for 16 weeks (Matthews et al., 1990; NTP, 1991b). First, the administered doses were duration adjusted to estimate an equivalent oral dose for animals exposed for seven rather than five days per week. Then, dichotomous models were used to fit dose-response data.

412

410

and response data used for the modeling are presented in Table 1-3. Table 1-3. Necrosis of the Neurons of the Hippocampus Selected for Dose-Response Modeling for TCEP from a 16-Week Study

EPA presents BMDLs based on BMRs of 5 and 10 percent ER from the best fit model. Based on the

severity of the endpoint and considering EPA's BMD Technical Guidance (U.S. EPA, 2012), EPA is

using the BMDL based on a BMR of 5 percent ER for this endpoint in the risk calculation. The doses

Dose (mg/kg/day)	Number of Animals	Incidence
0	10	0
16	10	0
31	10	0
63	10	0
125	10	8
250	10	10

413 414

424 425 426

427

423

The BMD modeling results for the necrosis of neurons in the hippocampus are summarized in Table 1-4. All models, except for the 1-degree Multistage model, provided an adequate fit (chi-square p-value > 0.1) to the data. Using a BMR of 10 percent extra risk, the BMDLs for the fit models were sufficiently close (differed by < 3-fold); therefore, the model with the lowest AIC (Probit) was selected. Using a BMR of 5 percent extra risk, however, BMDLs for the fit models differed by > 3-fold and the BMDS software recommended selection of the 2-degree Multistage model because it estimated the lowest BMDL. Although the 2-degree Multistage model provided overall adequate fit to the data, in the context of this data set, the high residuals at the key datapoints (-1.7 and 1.1) indicate a relatively poor fit in the key part of the dose-response curve. For this reason, the 2-degree Multistage was dropped from consideration. The BMDLs of the remaining models are sufficiently close (differed by < 3-fold), and the model with the lowest AIC (Probit) was selected.

Table 1-4. BMD Modeling Results for Necrosis of the Neurons of the Hippocampus in Female Rats in the 16-Week Study

Model	Goodness of Fit (Means)		BMD 5%ER	BMDL 5%ER	BMD 10%ER	BMDL 10%ER	Basis for Model
1/10001	p-value	AIC	(mg/kg- day)	(mg/kg- day)	(mg/kg- day)	(mg/kg- day)	Selection
Dichotomous Hill	1.00	14.01	98	52	102	61	
Gamma	0.99	14.52	69	49	76	57	The Probit
Log-Logistic	1.00	14.01	98	52	102	61	
Multistage 5	0.99	13.04	64	38	74	55	model is
Multistage 4	0.95	14.05	55	34	66	50	selected because of
Multistage 3	0.80	16.02	44	27	56	42	the lowest
Multistage 2	0.41	20.35	28	18	40	30	AIC.
Multistage 1	0.01	35.31	8	5	16	11	
Weibull	0.99	14.52	72	70	81	79	

Model	Goodness of Fit (Means)		BMD 5%ER	BMDL 5%ER	BMD 10%ER	BMDL 10%ER	Basis for Model
1/10001	p-value	AIC	(mg/kg- day)	(mg/kg- day)	(mg/kg- day)	(mg/kg- day)	Selection
Logistic	1.00	12.01	97	51	102	63	
Log-Probit	1.00	14.01	104	53	107	60	
Probit	1.00	12.01	90	50	96	61	

Plots of the Probit model with BMRs of 10 or 5 percent ER are shown in Figure 1-6 and Figure 1-7, respectively. Additional modeling details, including model parameters, goodness of fit at each dose, and log likelihood are shown below in Figure 1-8.

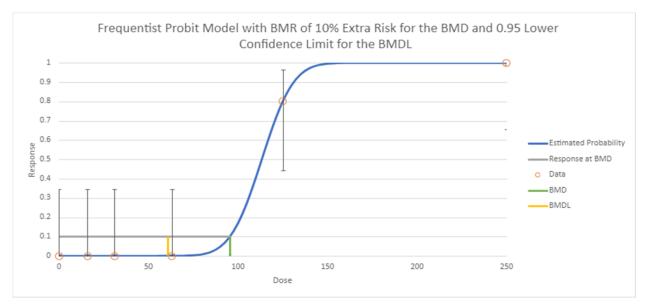


Figure 1-6. Plot of Response by Dose with Fitted Curve for the Selected Model (Probit) for Necrosis of the Neurons in the Hippocampus in Female Rats Exposed to TCEP Via Oral Gavage (16-Week Study) and BMR of 10 Percent Extra Risk

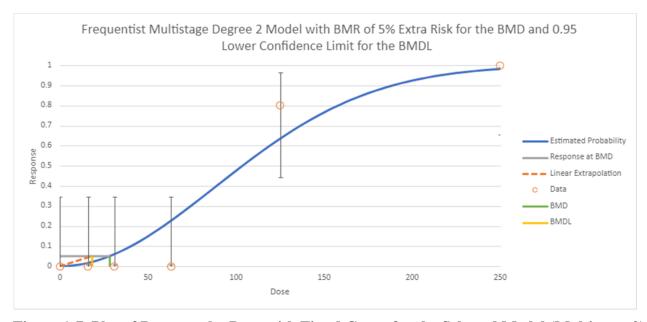


Figure 1-7. Plot of Response by Dose with Fitted Curve for the Selected Model (Multistage 2) for Necrosis of the Neurons in the Hippocampus in Female Rats Exposed to TCEP Via Oral Gavage (16-Week Study) and BMR of 5 Percent Extra Risk

1D					
MD	ark Dose				
	95.52742296	+			
MDL	60.85124313	-			
MDU	105.2988529	+			
c .	12.01096127	-			
value	0.99999996	+			
O.F.	5	1			
ni <sup>z</sup>	0.001459742	J			
Model P	arameters	1			
of Parameters	2				
Variable	Estimate				
a	-8.159526019				
ь	Bounded				
Goodne	ess of Fit	1			
Goodne	ess of Fit  Estimated Probability	Expected	Observed	Size	Scaled Residual
712	Estimated	Expected 1.68171E-15	Observed 0	Size	
Dose	Estimated Probability				Residual
Dose 0	Estimated Probability 1.68171E-16	1.68171E-15	0	10	Residual -4.1E-08
Dose 0 16	Estimated Probability 1.68171E-16 1.21284E-12	1.68171E-15 1.21284E-11	0	10	-4.1E-08 -3.48E-06
0 16 31	Estimated Probability 1.68171E-16 1.21284E-12 1.53766E-09	1.68171E-15 1.21284E-11 1.53766E-08	0 0 0	10 10 10	-4.1E-08 -3.48E-06 -0.000124

Figure 1-8. Details Regarding the Selected Model (Probit) for Necrosis of the Neurons in the Hippocampus in Female Rats Following Oral Exposure to TCEP in a 16-Week Chronic Toxicity Study

#### 1.1.1.3 Serum Cholinesterase Activity in Female Rats

Serum cholinesterase activity was decreased in female rats that were exposed to TCEP for 16 weeks (NTP, 1991b). First, the administered doses were duration adjusted to estimate an equivalent oral dose for animals exposed for seven rather than five days per week. Then, continuous models were used to fit dose-response data.

EPA modeled serum cholinesterase activity for BMRs of 1 SD and 10 percent RD according to EPA's *Benchmark Dose Technical Guidance* (U.S. EPA, 2012). The doses and response data used for the modeling are presented in Table 1-5.

Table 1-5. Decrease of Serum Cholinesterase Activity Selected for Dose-Response Modeling for TCEP from a 16-Week Study

Dose (mg/kg/day)	Number of Animals	Mean	SD
0	10	2064	354.18
16	8	1946	353.55
31	10	1808	332.04
63	10	1873	332.04
125	8	1550	294.16
250	5	1226	62.61

The BMD modeling results for serum cholinesterase activity are summarized in Table 1-6. The constant variance model did not provide adequate fit to the variance data, but the non-constant variance model did fit. With the non-constant variance model applied, all the models provided adequate fit to the means (test 4 p-value > 0.1). The BMDLs for the fit models were sufficiently close (differed by < 3-fold); therefore, the model with the lowest AIC was selected.

Table 1-6. Summary of BMD Modeling Results for Decreased of Serum Cholinesterase Activity in

Female Rats Following Oral Exposure to TCEP in a 16-Week Study

remaie Kats re	Goodness (Mean	s of Fit	BMD 1SD	BMDL 1SD	BMD 10%RD	BMDL 10%RD	Basis for Model
Model	Test 4 p-value	AIC	(mg/kg- day)	(mg/kg- day)	(mg/kg- day)	(mg/kg- day)	Selection
Exponential 2	0.634687	730	110.5	77.5	52.3	43.9	
Exponential 3	0.712137	730	147.3	84.7	87.0	45.9	The Linear
Exponential 4	0.634686	730	110.5	77.5	52.3	43.9	model is recommended
Exponential 5	0.503801	732	148.2	84.7	87.6	46.3	because it is the only model that
Hill	0.515392	732	147.4	83.0	82.8	41.1	provided adequate fit to
Polynomial Degree 5	0.538459	732	154.1	98.5	84.2	58.1	the means (test 4 p-value > 0.1).
Polynomial Degree 4	0.744042	730	153.7	98.5	84.4	57.3	The BMDLs for the fit models were
Polynomial Degree 3	0.744042	730	153.7	98.5	84.4	57.3	sufficiently close (differed by < 3-fold); therefore, the model with the lowest AIC was
Polynomial Degree 2	0.744042	730	153.7	98.5	84.4	57.3	
Power	0.726725	730	150.0	98.0	84.7	57.2	selected.
Linear	0.803824	729	129.6	96.3	64.3	56.8	

	Goodness of Fit (Means)		BMD 1SD	BMDL 1SD	BMD 10%RD	BMDL 10%RD	Basis for Model
Model	Test 4 p-value	AIC	(mg/kg- day)	(mg/kg- day)	(mg/kg- day)	(mg/kg- day)	Selection

<sup>&</sup>lt;sup>a</sup> Three significant figures

 Plots of the linear model with BMRs of one SD and 10 percent RD are shown in Figure 1-9 and Figure 1-10, respectively. Additional modeling details, including model parameters, goodness of fit at each dose, and log likelihood are shown below in Figure 1-11.



Figure 1-9. Plot of Response by Dose with Fitted Curve for the Selected Model (Linear) for Serum Cholinesterase Activity Decreases in Female Rats Exposed to TCEP Via Oral Gavage (16-Week Study) and BMR of 1SD (Non-constant Variance Assumed)

<sup>&</sup>lt;sup>b</sup> Selected model in bold; scaled residuals for selected model for doses 0, 16, 31, 63, 125, and 250 mg/kg-day were 0.555, 0.0167, -0.8154, 0.857, -0.7530, and 0.1978, respectively.

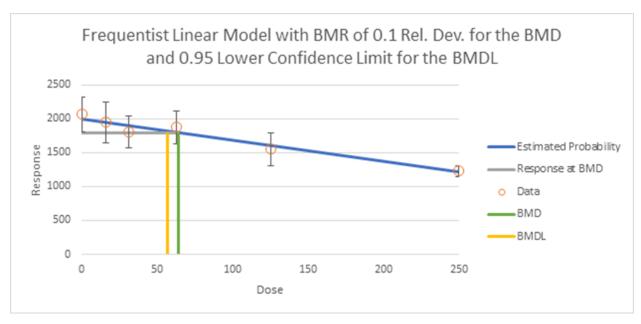


Figure 1-10. Plot of Response by Dose with Fitted Curve for the Selected Model (Linear) for Serum Cholinesterase Activity Decreases in Female Rats Exposed to TCEP Via Oral Gavage (16-Week Study) and BMR of 10 Percent Relative Deviation (Non-constant Variance Assumed)

Model Re	eulte									
WIOGEI IXE	Suits									
Be	enchma		1000000							
BMD			29.5518875							
BMDL		_	5.29342612							
BMDU			78.1793969							
AIC		-	28.5938025							
Test 4 P-valu	ie	0.	803823971							
D.O.F.			4							
Mo	del Par	ame	eters	8						
# of Parame	ters		4							
Variabl	e	1	Estimate							
g		19	993.420713							
beta1		-3	.101835334							
rho		6.	036517732							
Goodnes	ss of Fit		8							
Dose	Size		Estimated Median	Calc'd Median			Estimated SD	Salc'd SD	Observed SD	Scaled Residual
0	10		1993.420713	2064	206		401.848517		354.18	0.555411536
16	8		1943.791348	1946	194	16	372.404125	353.55	353.55	0.016774819
31	10		1897.263818	1808	180	8	346.143837	332.04	332.04	-0.815490402
63	10		1798.005087	1873	187	73	294.320203	332.04	332.04	0.805771181
125	8		1605.691296	1550	155	0	209.187846	294.16	294.16	-0.753001553
250	5		1217.96188	1226	122	26	90.836375	62.61	62.61	0.197869889
Likeli	ihoods	of Ir	iterest							
				# <u>of</u>			ALC			
Model A1		_	5 Likelihood* 63.3689517	Parame 7	reis	740	.737903			
A2	-		357.064324	12		_	.128648			
A3			59.4831033	8		_	.966207			
fitted			60.2969012	4		-	.593802			
R			76.3760886	2			.752177			
Te	ests of I	nter	est							
		-2*1	og(Likelihood		43		22			
Test			Ratio)	Test	f	-	value			
1		_	3.62352921	10		_	0.0001			
2			2.60925548	5		_	27329			
3			837558645	4			043747			
4		1	.62759584	4		0.80	382397			

Figure 1-11. Details Regarding the Selected Model (Linear) for Serum Cholinesterase Activity Decreases in Female Rats Following Oral Exposure to TCEP in a 16-Week Chronic Toxicity Study

#### 1.1.2 Reproductive and Developmental Toxicity

EPA modeled endpoints when one or more doses showed pairwise differences from controls and/or when a dose-response trend was evident in the data. EPA modeled litter data separately by sex as well as combined (males and females) as well as effects on male reproductive organs.

EPA did not present the BMD modeling results for several endpoints from NTP (1991a) that resulted in inadequate model fits. These endpoints included several for the F0 animals: cumulative days to litter (litter numbers 2 and 3); mean litters per pair; and live F1 pups per litter (both sexes and females). Also, although F1 fertility was modeled due to a statistically significant dose-response trend, the results are not presented because the BMD/BMDL ratio was greater than three and the BMDL was more than three times lower than the lowest dose tested. Testicular testosterone levels from Chen et al. (2015) were modeled but didn't fit any of the constant or non-constant variance models.

EPA also identified an anomaly in the data presented Table 4-4 within NTP (1991a) that affects the measures of sex of F2 pups born alive and live male F2 pups per litter (difference in proportion of males at 350 mg/kg-day). Therefore, although EPA modeled both effects (with an adequate model fit for live male F2 pups per litter), EPA is not presenting the results base on the identified error.

#### 1.1.2.1 Decreased Numbers of Seminiferous Tubules (Mice)

Chen et al. (2015) found decreases in numbers of seminiferous tubules in adolescent ICR mice after 35 days of exposure. Continuous models were used to fit data, and BMDLs based on BMRs of one SD and five percent RD from the best fit model are both presented. Based on the severity of the endpoint (considering EPA's BMD Technical Guidance (U.S. EPA, 2012)), EPA is using the BMDL based on a lower BMR (EPA used five percent RD) for this endpoint in the risk calculation. The doses and response data used for the modeling are presented in Table 1-7. There is uncertainty in using the BMDL based on a BMR of 5 percent because this BMR is lower than the responses observed in the study (decreases of 22.2 and 40.7 percent at 100 and 300 mg/kg-day, respectively).

Table 1-7. Decreased Numbers of Seminiferous Tubules in Mice and Associated Doses Selected for Dose-Response Modeling for TCEP

Dose (mg/kg-day)	Number of Mice	Mean	SD
0	7	24.3	5.29
100	7	18.9	3.17
300	7	14.4	2.65

Table 1-8 summarizes the BMD modeling results for decreased numbers of seminiferous tubules from Chen et al. (2015). The constant variance model provided adequate fit to the variance data and with this model applied, all models except the Exponential 4 and 5 models, provided adequate fit to the means (p-value > 0.1). BMDLs for the fit models were sufficiently close (< 3-fold difference). Therefore, EPA selected the model with the lowest Akaike information criterion (AIC). The software selected the Exponential 3 model, but EPA chose the Exponential 2 as the more parsimonious choice because Exponential 3 defaulted to the Exponential 2 model by bounding variable d at a value of one.

Table 1-8. Summary of BMD Modeling Results for Decreased Numbers of Seminiferous Tubules in Mice Following Oral Exposure to TCEP in a 35-Day Study (Constant Variance)<sup>a b</sup>

	Goodness	of Fit	BMD 1SD	BMDL 1SD	BMD 5%RD	BMDL 5%RD	Basis for Model	
Model	p-value	AIC	(mg/kg- day)	(mg/kg- day)	(mg/kg- day)	(mg/kg- day)	Selection	
Exponential 2	0.343	120	94.0	61.2	28.8	20.8	For the constant variance	
Exponential 3	0.343	120	94.0	61.2	28.8	20.9	model, all models except the Exponential 4 and 5	
Exponential 4	N/A	121	59.3	26.2	17.8	7.70	models, provided adequate fit to the means	
Exponential 5	< 0.0001	123	59.2	26.2	17.8	7.70	(p-value > 0.1). BMDLs were < 3-fold difference.	
Polynomial 2	0.223	121	118	82.7	37.1	29.0	EPA selected the Exponential 2, the model	
Power	0.223	121	118	82.7	37.1	29.0	with the lowest AIC (along with Exponential	

	Goodness of Fit BMD BMDL BMD 1SD 1SD 5%RD		BMDL 5%RD	Basis for Model			
Model	p-value	AIC	(mg/kg- day)	(mg/kg- day)	(mg/kg- day)	(mg/kg- day)	Selection
Linear	0.223	121	118	82.7	37.1	29.0	3). Exponential 2 is the more parsimonious.

<sup>&</sup>lt;sup>a</sup> Three significant figures

529

530531

532

533534

535536

Plots of the Exponential 2 model with BMRs of one SD and five percent RD are shown in Figure 1-12 and Figure 1-13, respectively. Additional modeling details, including model parameters, goodness of fit at each dose, and log likelihood are shown below in Figure 1-14.

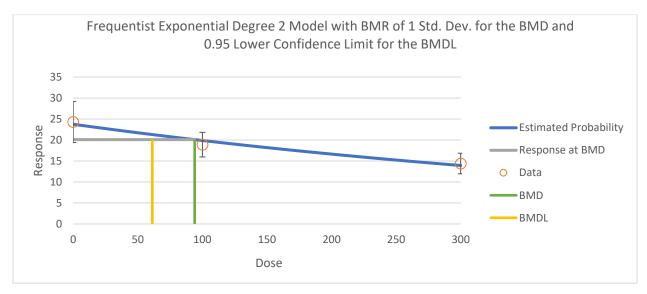


Figure 1-12. Plot of Response by Dose with Fitted Curve for the Selected Model (Exponential 2) for Decreased Numbers of Seminiferous Tubules in Mice Exposed to TCEP Via Oral Gavage in a 35-Day Study and BMR of 1SD (Constant Variance)

<sup>&</sup>lt;sup>b</sup> Selected model in bold; scaled residuals for selected model for doses 0, 100, and 300 mg/kg-day were 0.397, 0.711, and 0.338 respectively.



539540

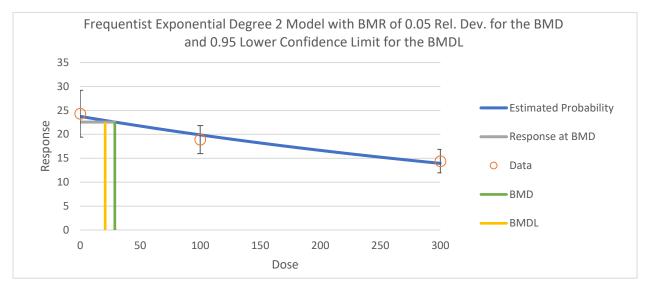


Figure 1-13. Plot of Response by Dose with Fitted Curve for the Selected Model (Exponential 2) for Decreased Numbers of Seminiferous Tubules in Mice Exposed to TCEP Via Oral Gavage in a 35-Day Study and BMR of 5 Percent Relative Deviation (Constant Variance)

Model Results	5											
Ben	ıchma	ark Dose										
BMD		94.01	164055									
BMDL		61.23	672499									
BMDU		177.5	203492									
AIC		120.0	453798									
Test 4 P-value	е	0.373	439526									
D.O.F.			1									
Mod	lel Pa	rameter	S									
# of Paramete	ers		3									
Variable		Est	imate									
a		23.75	164316									
b		0.001	778069			_						
Goodne	ss of	Fit										
Dose	,	Size	Estimat	ed	Calked	(	Observed	Es	stimated	Calc'd	Observed	Scaled
Бозе	,	3126	Media	n I	Median		Mean		SD	SD	SD	Residual
0		7	23.75164	316	24.3		24.3	3.6	55621204	5.29	5.29	0.396808451
100		7	19.88259	_	18.9		18.9		55621204	3.17	3.17	-0.711037874
300		7	13.93259	326	14.4		14.4	3.6	55621204	2.65	2.65	0.33823038
Likelih	oods	of Intere	est						•			
Model		Log Lik	elihood*	# of F	aramete	ers	AIC					
A1		-56.62	2659619		4		121.2531	92				
A2		-54.73	3789338		6		121.4757	87				
A3		-56.62	2659619		4		121.2531	92				
fitted		-57.02	2268989		3		120.0453	38				
R		-65.24	1552159	2			134.4910	43				
	Test	ts of Inte	rest						•			
Test		-2*Log	(Likelihood	ihood Ratio)		df	p-value	<u> </u>				
1			1.0152564		4		0.000314					
2			.77740563		2		0.151267					
3		3	.77740563	3	2		0.151267	79				
4		C	.79218738	37	1		0.373439	53				

Figure 1-14. Details Regarding the Selected Model (Exponential 2) for Decreased Numbers of Seminiferous Tubules in Mice in a 35-Day Study

#### **1.1.2.2** Decreases in Testes Weights (Mice)

Chen et al. (2015) identified decreased testes weights in adolescent ICR mice after 35 days exposure to TCEP. Continuous models were used to fit dose-response data. BMDLs based on BMRs of one SD and five percent RD from the best fit model are both presented. Based on EPA's BMD Technical Guidance (U.S. EPA, 2012), EPA is using the BMDL based on a BMR of one SD for this endpoint when comparing with other points of departure. The doses and response data used for modeling this endpoint are presented in Table 1-9.

Table 1-9. Decreased Testes Weights in Mice and Associated Doses Selected for Dose-Response Modeling for TCEP

Dose (mg/kg-day)	Number of Fertile Pairs	Mean	SD
0	7	0.32	0.053
100	7	0.28	0.04
300	7	0.27	0.019

Table 1-10 summarizes the BMD modeling results for decreased testes weights from Chen et al. (2015). The constant variance model did not provide adequate fit to the variance data, but the non-constant variance model did. With the non-constant variance model applied, all models except the Exponential 4 and 5 models provided adequate fit to the means. The BMDLs for the fit models were sufficiently close (< 3-fold difference). therefore, EPA chose the Exponential 3 model, the one with the lowest AIC was selected.

Table 1-10. Summary of BMD Modeling Results for Decreased Testes Weights in Mice Following Oral Exposure to TCEP in a 35-Day Study (Constant Variance)<sup>a b</sup>

	Goodnes	s of Fit	BMD 1SD	BMDL 1SD	BMD 5%RD	BMDL 5%RD	Basis for Model	
Model	p-value	AIC	(mg/kg- day)	(mg/kg- day)	(mg/kg- day)	(mg/kg- day)	Selection	
Exponential 2	0.659	-75.9	459	214	123	69.7	The non-constant	
<b>Exponential 3</b>	0.660	-75.9	467	214	125	69.7	variance model fit, and all models, except the	
Exponential 4	N/A	-73.9	469	34.8	125	0	Exponential 4 and 5, provided adequate fit to	
Exponential 5	65535	-72.0	_9999	0	81.4	0	the means. The BMDLs	
Polynomial 2	0.630	-75.8	460	224	131	77.0	for the fit models were < 3-fold different; EPA	
Power	0.630	-75.8	460	224	131	77.0	chose the model with the lowest AIC, the	
Linear	0.630	-75.8	460	225	131	77.0	Exponential 3.	

<sup>&</sup>lt;sup>a</sup> Three significant figures

Plots of the Exponential 3 model with BMRs of one SD and five percent RD are shown in Figure 1-15 and Figure 1-16, respectively. Additional modeling details, including model parameters, goodness of fit at each dose, and log likelihood are shown below in Figure 1-17.

<sup>&</sup>lt;sup>b</sup> Selected model in bold; scaled residuals for selected model for doses 0, 100, and 300 mg/kg-day were 0.778, 0.859, and 0.155 respectively.

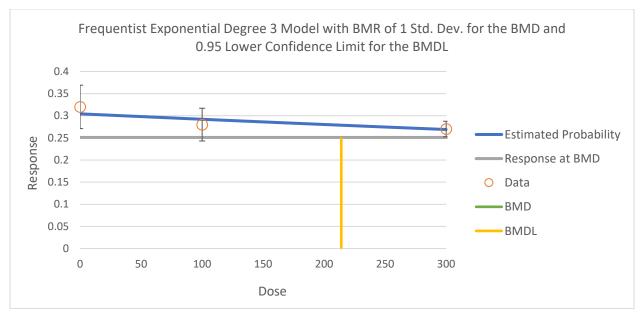


Figure 1-15. Plot of Response by Dose with Fitted Curve for the Selected Model (Exponential 3) for Decreased Testes Weights in Mice Exposed to TCEP Via Oral Gavage in a 35-Day Study and BMR of 1SD (Non-constant Variance)

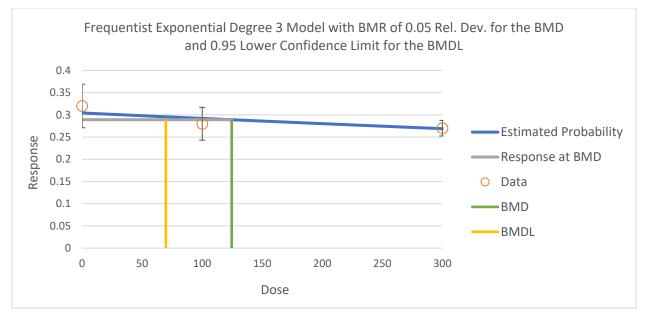


Figure 1-16. Plot of Response by Dose with Fitted Curve for the Selected Model (Exponential 3) for Decreased Testes Weights in Mice Exposed to TCEP Via Oral Gavage in a 35-Day Study and BMR of 5 Percent Relative Deviation (Non-constant Variance)

Model Results						
Benchma	ark Dose					
BMD	467.3440933					
BMDL	214.0626432					
BMDU	792.6585915					
AIC	-75.85266567					
Test 4 P-value	0.660168789					
D.O.F.	1					

Model Parameters						
# of Parameters	5					
Variable	Estimate					
a	0.304323078					
b	0.000411915					
d	Bounded					
rho	17.60378098					

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Goodnes	s of Fit							
Dose	Size	Estimated Median	Calc'd Median	Observed Mean	Estimated SD	Calc'd SD	Observed SD	Scaled Residual
0	7	0.304323078	0.32	0.32	0.05329016	0.053	0.053	0.778328256
100	7	0.292042232	0.28	0.28	0.03708412	0.04	0.04	-0.859148124
300	7	0.268947308	0.27	0.27	0.01795849	0.019	0.019	0.155088856

Likelihoods	s of Interest		
Model	Log Likelihood*	# of Parameters	AIC
A1	39.4830962	4	-70.966192
A2	42.65846137	6	-73.316923
A3	42.02299215	5	-74.045984
fitted	41.92633284	4	-75.852666
R	36.39113618	2	-68.782272

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Test	-2*Log(Likelihood Ratio)	Test df	p-value
1	12.53465039	4	0.01378827
2	6.350730343	2	0.04177884
3	1.270938445	1	0.2595907
4	0.193318625	1	0.66016879

Figure 1-17. Details Regarding the Selected Model (Exponential 3) for Decreased Testes Weights in Mice in a 35-Day Study

#### 1.1.2.3 Live Male F1 Pups per Litter (Mice)

NTP (1991a) identified decreases in the number of live male F1 mouse pups per litter. BMDLs based on BMRs of one SD and five percent RD from the best fit model are both presented. Based on the severity of the endpoint that was observed in offspring and considering EPA's BMD Technical Guidance (U.S.

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<sup>a</sup> Three significant figures

doses and response data used for the modeling are presented in Table 1-11. Table 1-11. F1 Live Male F1 Pups per Litter in Mice and Associated Doses

**Selected for Dose-Response Modeling for TCEP** 

Dose (mg/kg-day)	Number of Fertile Pairs	Mean	SD
0	37	6.4	1.82
175	18	6.1	1.27
350	18	5.1	1.7
700	18	3.9	1.27

EPA, 2012), EPA is using the BMDL based on a BMR of five percent RD for this endpoint when

comparing with other points of departure. Continuous models were used to fit dose-response data. The

Table 1-12 summarizes the BMD modeling results for live male F1 mice per litter from NTP (1991a). The constant variance model provided an adequate fit to the variance data. With the constant variance model applied, all models, except for the Exponential 5 and Hill models, provided adequate fit to the means. The BMDLs for the fit models were sufficiently close (differed by < 3-fold). The 2-degree and 3-degree Polynomial models converged on the same model and had the lowest AIC. EPA chose the 2degree Polynomial model because it had the lowest AIC and was the more parsimonious choice.

Table 1-12. Summary of BMD Modeling Results for Live Male F1 Pups per Litter in Mice Following Oral Exposure to TCEP in a Continuous Breeding Study (Constant Variance)<sup>a b</sup>

Model	Goodness of Fit		SOLFIL		BMD 5%RD	BMDL 5%RD	Basis for Model
Model	p-value	AIC	(mg/kg- day)	(mg/kg- day)	(mg/kg- day)	(mg/kg- day)	Selection
Exponential 2	0.583	347	402	286	74.3	56.1	With constant variance option, all
Exponential 3	0.529	348	447	298	125	58.3	models (except Exponential 5 and
Exponential 4	0.583	347	402	286	74.3	56.1	Hill) provided adequate fit to the
Exponential 5	N/A	350	393	281	180	59.5	means and had BMDLs that were
Hill	N/A	350	398	275	180	50.6	sufficiently close (< 3-fold difference).
Polynomial 3	0.747	346	455	330	103	71.5	The 2-degree and 3-degree Polynomial models converged and had the lowest
Polynomial 2	0.747	346	455	330	103	71.5	
Power	0.475	348	457	331	115	71.7	AIC. EPA chose the 2-degree Polynomial
Linear	0.717	347	431	329	88.7	71.3	model as most parsimonious.

<sup>&</sup>lt;sup>b</sup> Selected model in bold; scaled residuals for selected model for doses 0, 175, 350, and 700 mg/kg-day were 0.155, 0.594, 0.446, and 0.0743, respectively.

Plots of the Polynomial 2 model with BMRs of one SD and five percent RD are shown in Figure 1-18 and Figure 1-19, respectively. Additional modeling details, including model parameters, goodness of fit at each dose, and log likelihood are shown below in Figure 1-20.

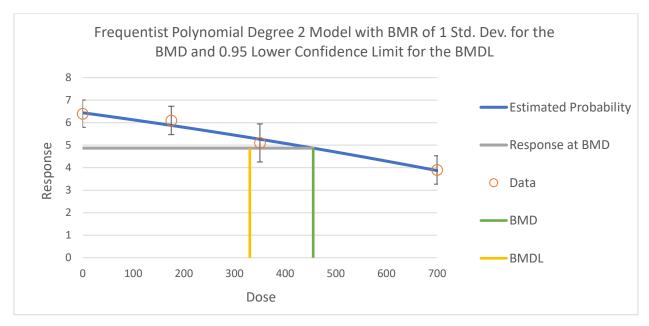


Figure 1-18. Plot of Response by Dose with Fitted Curve for the Selected Model (Polynomial 2) for Live Male F1 Pups per Litter in Mice Exposed to TCEP Via Oral Gavage in a Continuous Breeding Study and BMR of 1SD (Constant Variance)

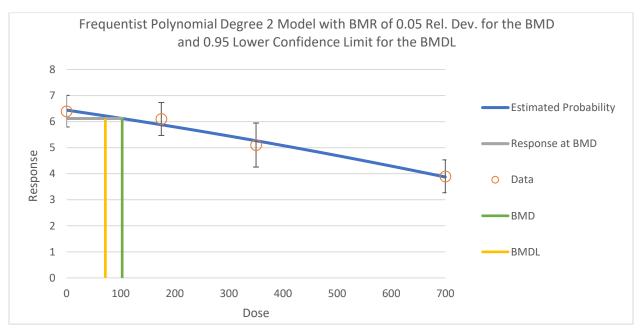


Figure 1-19. Plot of Response by Dose with Fitted Curve for the Selected Model (Polynomial 2) for Live Male F1 Pups per Litter in Mice Exposed to TCEP Via Oral Gavage in a Continuous Breeding Study and BMR of 5 Percent Relative Deviation (Constant Variance)

#### **Model Results**

Benchmark Dose					
BMD	455.3158283				
BMDL	330.1312755				
BMDU	636.807465				
AIC	346.4826916				
Test 4 P-value	0.746854679				
D.O.F.	2				

Model Parameters					
# of Parameters	4				
Variable	Estimate				
g	6.440162141				
beta1	-0.003046379				
beta2	Bounded				

Goodness	of Fit
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Dose	Size	Estimated Median	Calc'd Median	Observed Mean	Estimated SD	Calc'd SD	Observed SD	Scaled Residual
0	37	6.440162141	6.4	6.4	1.57120085	1.82	1.82	-0.155484111
175	18	5.879844619	6.1	6.1	1.57120085	1.27	1.27	0.59447535
350	18	5.265124542	5.1	5.1	1.57120085	1.7	1.7	-0.445878131
700	18	3.872476719	3.9	3.9	1.57120085	1.27	1.27	0.074319837

Likelihoods	of Interest		
Model	Model Log Likelihood*		AIC
A1	-169.9494612	5	349.898922
A2	-167.3861158	8	350.772232
A3	-169.9494612	5	349.898922
fitted	-170.2413458	3	346.482692
R	-184.5846567	2	373.169313

Tes			
Test	-2*Log(Likelihood Ratio)	Test df	p-value
1	34.39708179	6	< 0.0001
2	5.126690709	3	0.16275184
3	5.126690709	3	0.16275184
4	0.583769305	2	0.74685468

Figure 1-20. Details Regarding the Selected Model (Polynomial 2) for Live Male F1 Pups per Litter in a Continuous Breeding Study

#### 1.1.2.4 Live F2 Pups per Litter (Mice)

NTP (1991a) identified decreased mean numbers of F2 mice pups per litter in the F2 generation. Continuous models were used to fit dose-response data. BMDLs based on BMRs of one SD and five percent RD from the best fit model are both presented. Based on the severity of this effect in offspring and considering EPA's BMD Technical Guidance (U.S. EPA, 2012), EPA is using the BMDL based on

a BMR of five percent RD for this endpoint when comparing with other points of departure. The doses and response data used for the modeling are presented in Table 1-13.

Table 1-13. Live F2 Pups per Litter in Mice and Associated Doses Selected for Dose-Response Modeling for TCEP

Dose (mg/kg-day)	Number of Fertile Pairs	Mean	SD
0	17	11.4	2.06
175	18	11	2.12
350	14	7.6	4.12

Table 1-14 summarizes the BMD modeling results for live pups per litter from NTP (1991a). The constant variance model did not provide adequate fit to the variance data, but the non-constant variance model did provide an adequate fit. Applying the non-constant variance model, only the 2-degree Polynomial provided adequate fit to the means (test 4 p-value > 0.1); therefore, this model was selected.

Table 1-14. Summary of BMD Modeling Results for Live F2 Pups per Litter in Mice Following Oral Exposure to TCEP in a Continuous Breeding Study (Non-constant Variance)<sup>a</sup> b

Model	Goodness of Fit	BMD 1SD	BMDL 1SD	BMD 5%RD	BMDL 5%RD	Basis for Model	
	p-value	AIC	(mg/kg- day)	(mg/kg- day)	(mg/kg- day)	(mg/kg- day)	Selection
Exponential 2	0.0157	241	230	133	67.3	40.9	
Exponential 3	N/A	237	284	203	198	102	Of the non-constant variance models (the only ones that adequately fit the variance data), the 2-degree Polynomial provided adequate fit to the means (test 4 p-value > 0.1) and EPA selected this model.
Exponential 4	0.0157	241	230	133	67.4	40.9	
Exponential 5	N/A	237	284	203	198	102	
Hill	< 0.0001	239	223	180	185	155	
Polynomial 2	0.335	236	252	192	139	76.5	
Power	N/A	238	343	199	326	301	
Linear	0.0232	240	223	140	69.3	45.7	

<sup>&</sup>lt;sup>a</sup> Three significant figures

Plots of the Polynomial 2 model with BMRs of one SD and five percent RD are shown in Figure 1-21 and Figure 1-22, respectively. Additional modeling details, including model parameters, goodness of fit at each dose, and log likelihood are shown below in Figure 1-23.

<sup>&</sup>lt;sup>b</sup> Selected model in bold; scaled residuals for selected model for doses 0, 175, and 350 mg/kg-day were 0.418, 0.624, and 0.293, respectively.

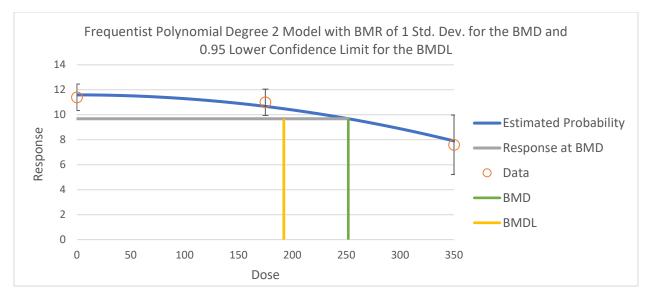


Figure 1-21. Plot of Response by Dose with Fitted Curve for the Selected Model (Polynomial 2) for Live F2 Pups per Litter in Mice Exposed to TCEP Via Oral Gavage in a Continuous Breeding Study and BMR of 1SD (Non-constant Variance)

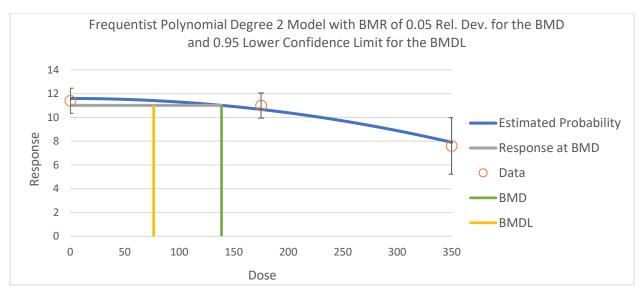


Figure 1-22. Plot of Response by Dose with Fitted Curve for the Selected Model (Polynomial 2) for Live F2 Pups per Litter in Mice Exposed to TCEP Via Oral Gavage in a Continuous Breeding Study and BMR of 5 Percent Relative Deviation (Non-constant Variance)

#### **Model Results**

Benchmark Dose (1 SD)				
BMD	251.9403458			
BMDL	192.0823998			
BMDU	367.1866626			
AIC	236.1697373			
Test 4 P-value	0.334975944			
D.O.F.	1			

Model Parameters				
# of Parameters	5			
Variable	Estimate			
g	11.5935616			
beta1	Bounded			
beta2	-3.01002E-05			
rho	-3.746949647			

Goodness of Fit								
Dose	Size	Estimated Median	Calc'd Median	Observed Mean	Estimated SD	Calc'd SD	Observed SD	Scaled Residual
0	17	11.5935616	11.4	11.4	1.91057668	2.06	2.06	-0.417714158
175	18	10.67174339	11	11	2.23138704	2.12	2.12	0.624129657
350	14	7.906288771	7.6	7.6	3.91398177	4.12	4.12	-0.292803521

Likelihoo	ds of Interest		
Model	Log Likelihood*	# of Parameters	AIC
A1	-118.7238354	4	245.447671
A2	-113.6129329	6	239.225866
A3	-113.6200861	5	237.240172
fitted	-114.0848686	4	236.169737
R	-126.2176267	2	256.435253

T			
Test	-2*Log(Likelihood Ratio)	Test df	p-value
1	25.20938769	4	< 0.0001
2	10.2218049	2	0.00603064
3	0.014306312	1	0.90479289
4	0.929565159	1	0.33497594

Figure 1-23. Details Regarding the Selected Model (Polynomial 2) for Live F2 Pups per Litter in Mice in a Continuous Breeding Study

#### 1.1.2.5 F0 Fertility in Mice

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NTP (1991a) identified increases in the number of non-fertile pairs per number of cohabiting mice for litter five from the F0 generation. Dichotomous models were fit to the incidence data. EPA chose a BMR of five percent ER according to EPA's *BMD Technical Guidance* (U.S. EPA, 2012) to compare

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Table 1-15. F0 Non-fertility in Mice and Associated Doses Selected for **Dose-Response Modeling for TCEP** 

Dose (mg/kg-day)	Number of Animals Cohabiting	Incidence of Nonfertility
0	38	3
175	19	2
350	18	5
700	18	18

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Table 1-16 summarizes the BMD modeling results for F0 non-fertile mice from NTP (1991a). The Dichotomous Hill, Gamma, Log-logistic, 3-degree Multistage, Weibull, and Log-probit models provided an adequate fit (chi-square p-value > 0.1) to the data. The BMDLs for the fit models were sufficiently close (differed by < 3-fold). Therefore, EPA chose the model with the lowest AIC.

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Table 1-16. Summary of BMD Modeling Results for F0 Non-fertile Mice Following Oral Exposure

to TCEP in a Continuous Breeding Study<sup>a b</sup>

M 11	Goodnes	ss of Fit	BMD	BMDL	D . C M LIGH .
Model	p-value AIC (mg/kg/day) (mg/kg/day)		Basis for Model Selection		
Dichotomous Hill	0.947	59.2	320	225	
Gamma	0.863	59.5	275	200	The Dichotomous Hill,
Log-Logistic	0.947	59.2	320	225	Gamma, Log-logistic, 3-
Multistage 3	0.456	61.3	175	82.4	degree Multistage, Weibull, and Log-probit models
Multistage 2	0.0697	66.5	108	65.9	provided adequate fits to the data (chi-square p-value >
Multistage 1	0.000911	78.7	29.6	20.7	0.1). The BMDLs for the fit models were sufficiently
Weibull	0.773	61.1	271	161	close (differed by < 3-fold); therefore, EPA selected the
Logistic	0.0878	64.8	108	72.9	model with the lowest AIC.
Log-Probit	0.741	61.2	329	229	
Probit	0.0609	65.6	90.6	62.5	

<sup>&</sup>lt;sup>a</sup> Three significant figures

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Figure 1-24 shows the log-logistic model, the chosen model for F0 fertility with a BMR of five percent RD. shows additional modeling details, including model parameters, goodness of fit at each dose, and log likelihood.

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<sup>&</sup>lt;sup>b</sup> Selected model in bold; scaled residuals for selected model for doses 0, 175, 350, and 700 mg/kg-day were -0.191, 0.270, -1.7E-04, and 1.54E-2, respectively.

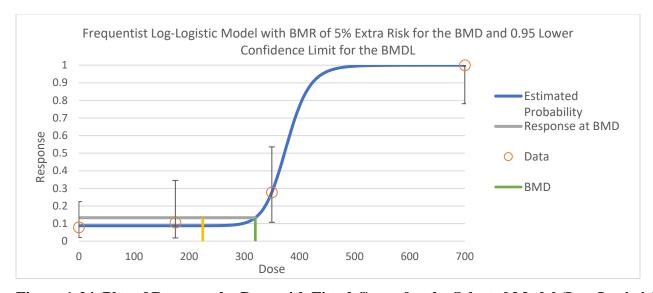


Figure 1-24. Plot of Response by Dose with Fitted Curve for the Selected Model (Log-Logistic) for F0 Non-fertile Mice Exposed to TCEP Via Oral Gavage in a Continuous Breeding Study and BMR of 5 Percent Extra Risk

#### **Model Results**

Benchmark Dose				
BMD	320.0613905			
BMDL	224.693377			
BMDU	362.8647716			
AIC	59.15490795			
P-value	0.946557537			
D.O.F.	2			
Chi <sup>2</sup>	0.109847042			

Model Parameters				
# of Parameters	3			
Variable	Estimate			
G	0.08771758			
A	-106.7776698			
В	Bounded			

Goodness	of Fit				
Dose	Estimated Probability	Expected	Observed	Size	Scaled Residual
0	0.08771758	3.333268046	3	38	-0.191115
175	0.087718496	1.666651425	2	19	0.2703409
350	0.277795687	5.000322359	5	18	-0.00017
700	0.999986778	17.999762	18	18	0.0154275

Analysis of Deviance					
Model	Log Likelihood	# of Parameters	Deviance	Test d.f.	P Value
Full Model	-27.52383488	4	-	-	N/A
Fitted Model	-27.57745398	2	0.1072382	2	0.9477931
Reduced Model	-56.89485404	1	58.7420383	3	< 0.0001

Figure 1-25. Details Regarding the Selected Model (Log-Logistic) for F0 Non-fertile Mice in a Continuous Breeding Study

## 1.1.3 Kidney Toxicity

 EPA selected multiple kidney endpoints for quantitative dose-response analysis with BMDS, including histopathological lesions and kidney weights. EPA modeled kidney weight changes when a pairwise change from controls and/or dose-response trend was evident in the data (*e.g.*, a statistically significant change was identified). The best measures are kidney weight changes relative to body weight (to account for any changes that are primarily related to body weight changes). EPA presents the female rat relative kidney weight data from the 16-week NTP (1991b) study after dropping the highest dose from the models and considers this to be appropriate due to the decreased survival at the highest dose (5 of 10 animals died). However, EPA could not model the female rat absolute kidney without dropping the *two* highest doses and therefore, EPA is not presenting these data. In the 16-week study (NTP, 1991b), male kidney weights were increased only at the highest doses and therefore, EPA did not conduct BMD modeling for these changes.

## 1.1.3.1 Renal Tubule Hyperplasia in Male Rats

There was an increased incidence of renal tubule hyperplasia in male rats exposed to TCEP for two years in a chronic toxicity study by NTP (1991b). First, the administered doses were duration adjusted to estimate an oral dose for animals exposed for seven days per week rather than five days per week. Then, dichotomous models were used to fit dose-response data.

A BMR of 10 percent ER was chosen according to *BMD Technical Guidance* (<u>U.S. EPA, 2012</u>). The doses and response data used for the modeling are presented in Table 1-17.

Table 1-17. Incidence of Renal Tubule Hyperplasia Selected for Dose-Response Modeling for TCEP

1120 070 11119 1 0 1 1 0 1 1 1		
Dose (mg/kg/day)	Number of Animals	Incidence
0	50	0
31	50	2
63	50	24

The BMD modeling results for renal tubule hyperplasia are summarized in Table 1-18. The best fitting model was the Gamma based on the AIC (lower values indicate a better fit), chi-square goodness of fit p-value (higher value indicates a better fit) and visual inspection. A plot of the model is shown in Figure 1-26. The model version number, model form, benchmark dose calculation, parameter estimates, and estimated values are shown below in Figure 1-27.

Table 1-18. Summary of BMD Modeling Results for Renal Tubule Hyperplasia in Male Rats Following Oral Exposure to TCEP in a Two-Year Chronic Study<sup>a</sup>

Madal	Goodness of fit		oodness of fit BMD BMDL		D . C M LIGH &
Model	p-value AIC (mg/kg/day) (mg/kg/day)	Basis for Model Selection			
Dichotomous Hill	N/A	92.0	38.6	30.9	
Gamma	0.999	90.0	38.3	30.7	
Log-Logistic	N/A	92.0	38.8	30.7	The Gamma, Logistic, and
Multistage 2	0.0452	95.0	28.0	22.4	Probit models provided adequate fit to the data (chi-
Multistage 1	0.000826	104	15.6	11.5	square p-value > 0.1). The BMDLs were sufficiently
Weibull	N/A	92.0	39.5	30.7	close (differed by < 3-fold); therefore, the model with the
Logistic	0.749	90.2	41.9	34.2	lowest AIC was selected.
Log-Probit	N/A	94.8	53.7	27.0	
Probit	0.896	90.1	40.0	32.4	

<sup>&</sup>lt;sup>a</sup> Selected model in bold; scaled residuals for selected model for doses 0, 31, 63 were −8.73E−04, 1.40E−05, 4.49E−05, respectively.

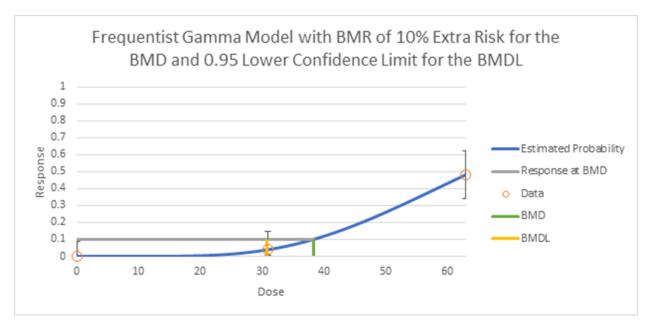


Figure 1-26. Plot of Mean Response by Dose with Fitted Curve for the Selected Model (Gamma) for Renal Tubule Hyperplasia in Male Rats Exposed to TCEP Via Oral Gavage in mg/kg/day; BMR 10 Percent Extra Risk

Ben	chmark Dose
BMD	38.3027844
BMDL	30.70367203
BMDU	44.21349368
AIC	90.02911301
P-value	0.999302723
D.O.F.	1
Chi <sup>2</sup>	7.63714E-07

Model Parameters					
# <u>of</u> Parameters	3				
Variable	Estimate				
g	Bounded				
a	7.529213517				
b	0.112149316				

Goodness	of Fit				
Dose	Estimated Probability	Expected	Observed	Size	Scaled Residual
0	1.523E-08	7.61499E-07	0	50	-0.000873
31	0.039999613	1.999980646	2	50	1.397E-05
63	0.479996825	23.99984124	24	50	4.494E-05

Analysis of Deviance					
Model	Log Likelihood	# of Parameters	Deviance	Test d.f.	P Value
Full Model	-43.01455574	3	-	-	NA
Fitted Model	-43.0145565	2	1.5252E-06	1	0.9990146
Reduced Model	-69.16986999	1	52.3106285	2	<0.0001

Figure 1-27. Details Regarding the Selected Model (Gamma) for Renal Tubule Hyperplasia in Male Rats Following Oral Exposure to TCEP in a Two-Year Chronic Toxicity Study

## 1.1.3.2 Renal Tubule Hyperplasia in Female Rats

There was an increased incidence of renal tubule hyperplasia in female rats exposed to TCEP for two years in a chronic toxicity study by NTP (1991b). First, the administered doses were duration adjusted to estimate an equivalent oral dose for animals exposed for seven days per week rather than five days per week. Then, dichotomous models were used to fit dose-response data.

A BMR of 10 percent extra risk was chosen according to *BMD Technical Guidance* (<u>U.S. EPA, 2012</u>). The doses and response data used for the modeling are presented in Table 1-19.

Table 1-19. Incidence of Renal Tubule Hyperplasia Selected for Dose-Response Modeling for TCEP

Dose (mg/kg/day)	Number of Animals	Incidence
0	50	0
31	50	3
63	50	16

The BMD modeling results for renal tubule hyperplasia are summarized in Table 1-20. The best fitting model was the Multistage Degree 2 based on the AIC (lower values indicate a better fit), chi-square goodness of fit p-value (a higher value indicates a better fit) and visual inspection. A plot of the model is shown in Figure 1-28. The model version number, model form, benchmark dose calculation, parameter estimates, and estimated values are shown below in Figure 1-29.

Table 1-20. Summary of BMD Modeling Results for Renal Tubule Hyperplasia in Female Rats Following Oral Exposure to TCEP in a Two-Year Chronic Study<sup>a</sup>

M 11	Goodne	ss of Fit	PMD PMDI		D ' C M LIGI C		
Model	p-value	AIC	BMD	BMDL	Basis for Model Selection		
Dichotomous Hill	N/A	91.4	37.1	25.5			
Gamma	N/A	91.4	37.6	25.3			
Log-Logistic	0.999	89.4	37.7	25.5	The Log-logistic, Multistage 1- and 2-degree, Logistic, and		
Multistage 2	0.804	87.9	34.2	23.2	Probit models provided adequate fit to the data (chi-		
Multistage 1	0.170	91.4	22.9	16.0	square p-value > 0.1). The BMDLs were sufficiently close		
Weibull	N/A	91.4	38.1	25.2	(differed by < 3-fold); therefore, the model with the		
Logistic	0.509	90.1	42.9	35.3	lowest AIC was selected.		
Log-Probit	N/A	95.6	56.9	12.5			
Probit	0.642	89.7	40.8	33.2			

<sup>&</sup>lt;sup>a</sup> Selected model in bold; scaled residuals for selected model for doses 0, 31, 63 were −8.73E−04, −0.584, and 0.308, respectively.

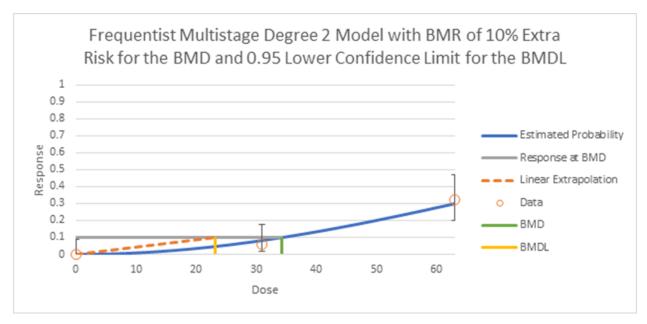


Figure 1-28. Plot of Mean Response by Dose with Fitted Curve for the Selected Model (Multistage Degree 2) for Renal Tubule Hyperplasia in Female Rats Exposed to TCEP Via Oral Gavage in mg/kg/day; BMR 10 Percent Extra Risk

Model Results					
Benchn	nark Dose	]			
BMD	34.23865288	1			
BMDL	23.23697912	1			
BMDU	41.90302307	1			
AIC	87.85147946	1			
P-value	0.80421339	1			
D.O.F.	2	]			
Chi <sup>2</sup>	0.435781269	]			
Slope Factor	0.004303485	]			
		1			
	arameters	-			
# <u>of</u> Parameters	3				
Variable	Estimate	1			
g	Bounded	-			
b1	Bounded	1			
b2	8.98762E-05				
Canda	ess of Fit	1			
Goodn	Estimated				Scaled
Dose	Probability	Expected	Observed	Size	Residual
0	1.523E-08	7.61499E-07	0	50	-0.000873
31	0.082746138	4.13730689	3	50	-0.583813
63	0.300030485	15.00152425	16	50	0.3081274
Analysis of Deviance					
Model	Log Likelihood	# of Parameters	Deviance	Test d.f.	P Value
Full Model	-42.691849	3	-	-	NA
Fire and Administration	-42.92573973	1	0.46778145	2	0.7914483
Fitted Model	-42.92373973	1	0.40770143	_	0.7311103

Figure 1-29. Details Regarding the Selected Model (Gamma) for Renal Tubule Hyperplasia in Female Rats Following Oral Exposure to TCEP in a Two-Year Chronic Toxicity Study

## 1.1.3.3 Renal Tubule Karyomegaly in Male Mice

There was an increased incidence of renal tubule karyomegaly (nuclear enlargement) in male mice exposed to TCEP for two years in a chronic toxicity study by NTP (1991b). First, the administered doses were duration adjusted to estimate an equivalent oral dose for animals exposed for seven days per week rather than five days per week. Then, dichotomous models were used to fit dose-response data.

A BMR of ten percent ER was chosen according to *BMD Technical Guidance* (U.S. EPA, 2012). The doses and response data used for the modeling are presented in Table 1-21.

Table 1-21. Incidence of Renal Tubule Karyomegaly Selected for Dose-Response Modeling for TCEP

Dose (mg/kg-day)	Number of Animals	Incidence
0	50	2
125	50	16
250	50	39

The BMD modeling results for renal tubule karyomegaly (nuclear enlargement) are summarized in Table 1-22. The best fitting model was the Probit based on the AIC (lower values indicate a better fit), chi-square goodness of fit p-value (higher value indicates a better fit) and visual inspection. A plot of the Probit model is shown in Figure 1-30. The model version number, model form, benchmark dose calculation, parameter estimates, and estimated values are shown below in Figure 1-31.

Table 1-22. Summary of BMD Modeling Results for Renal Tubule Karyomegaly (Nuclear Enlargement) in Male Mice Following Oral Exposure to TCEP in a Two-Year Chronic Toxicity Study<sup>a</sup>

Study					
M-J-1	Goodness of Fit		BMD	BMDL	Dania fan Madal Calastian
Model	p-value	AIC	(mg/kg-day)	(mg/kg-day)	Basis for Model Selection
Dichotomous Hill	65,535	140	81.6	50.8	
Gamma	N/A	138	77.9	42.3	
Log-Logistic	N/A	138	81.1	50.8	The Multistage 2-degree,
Multistage 2	0.846	136	67.5	32.0	Logistic, and Probit models provided adequate fit to the data
Multistage 1	0.0194	142	23.7	18.7	(chi-square p-value > 0.1). The BMDLs were sufficiently close
Weibull	N/A	138	71.0	38.6	(differed by <3-fold); therefore, the model with the lowest AIC
Logistic	0.686	136	69.4	54.5	was selected.
Log-Probit	N/A	153	224	0	
Probit	0.935	136	63.9	50.5	

<sup>&</sup>lt;sup>a</sup> Selected model in bold; scaled residuals for selected model for doses 0, 125, 250, were −0.0493, 0.0573, and 0.0307, respectively.

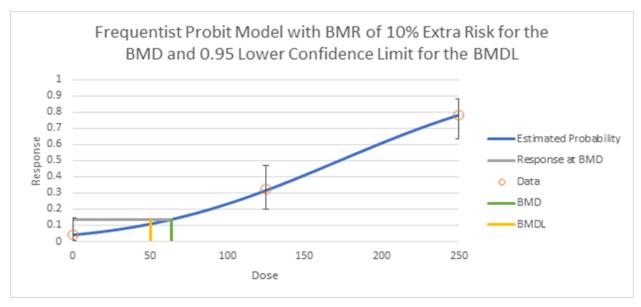


Figure 1-30. Plot of Mean Response by Dose with Fitted Curve for the Selected Model (Probit) for Renal Tubule Karyomegaly (Nuclear Enlargement) in Male Mice Exposed to TCEP Via Oral Gavage in mg/kg/day; BMR 10 Percent Extra Risk

Model Results					
Benchma	ork Dasa				
BMD	1				
	63.86977148	-			
BMDL	50.49773419	4			
BMDU	80.89809675	-			
AIC	136.1788322	-			
P-value	0.934968193				
D.O.F.	1				
Chi²	0.006657864				
Model Pa					
# of Parameters	rameters 2	,			
Variable	Estimate				
а	-1.734794541				
b	0.010052255				
~	0.020002200	_			
Goodne	ss of Fit				
Dose	Estimated	Eurostad		6:	Scaled
Dose	Probability	Expected	Observed	Size	Residual
0	0.041388602	2.069430088	2	50	-0.049295
125	0.31623162	15.81158102	16	50	0.0573037
250	0.781794818	39.0897409	39	50	-0.030727
	'				
Analysis of Deviance					
Model	Log Likelihood	# of Parameters	Deviance	Test d.f.	P Value
Full Model	-66.08607834	3	-	-	NA
Fitted Model	-66.08941609	2	0.0066755	1	0.9348823

Figure 1-31. Details Regarding the Selected Model (Probit) for Renal Tubule Karyomegaly (Nuclear Enlargement) in Male Mice Following Oral Exposure to TCEP in a Two-Year Chronic Toxicity Study

## 1.1.3.4 Renal Tubule Karyomegaly in Female Mice

There was an increased incidence of renal tubule karyomegaly (nuclear enlargement) in female mice exposed to TCEP for 2 years in a chronic toxicity study by NTP (1991b). First, the administered doses were duration adjusted to estimate an equivalent oral dose for animals exposed for seven days per week rather than five days per week. Then, dichotomous models were used to fit dose-response data.

A BMR of 10 percent ER was chosen according to *BMD Technical Guidance* (<u>U.S. EPA, 2012</u>). The doses and response data used for the modeling are presented in Table 1-23.

Table 1-23. Incidence of Renal Tubule Karyomegaly Selected for Dose-Response Modeling for TCEP

Dose (mg/kg/day)	Number of Animals	Incidence
0	50	0
125	49	5
250	50	44

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The BMD modeling results for renal tubule karyomegaly (nuclear enlargement) are summarized in Table 1-24. The best fitting model was the Gamma based on the AIC (lower values indicate a better fit), chi-square goodness of fit p-value (higher value indicates a better fit) and visual inspection. A plot of the Gamma model is shown in Figure 1-32. The model version number, model form, benchmark dose calculation, parameter estimates, and estimated values are shown below in Figure 1-33.

Table 1-24. Summary of BMD Modeling Results for Renal Tubule Karyomegaly (nuclear enlargement) in Female Mice Following Oral Exposure to TCEP in a Two-Year Chronic Toxicity Study<sup>a</sup>

M. J.I	Goodness of Fit		BMD	BMDL	Basis for Model
Model	p-value	$AIC^c$	(mg/kg/day)	(mg/kg/day)	Selection
Dichotomous Hill	N/A	75.0	125	108	
Gamma	0.999	73.0	125	107	The Gamma, Logistic,
Log-Logistic	N/A	75.0	125	108	and Probit models provided adequate fit
Multistage 2	0.00143	86.7	68.1	57.1	to the data (chi-square p-value > 0.1). The
Multistage 1	< 0.0001	110	25.5	20.1	BMDLs were
Weibull	N/A	75.0	124	102	sufficiently close (differed by < 3-fold);
Logistic	0.767	73.2	126	103	therefore, the model with the lowest AIC
Log-Probit	N/A	75.0	125	109	was selected.
Probit	0.943	73.0	125	102	

<sup>&</sup>lt;sup>a</sup> Three significant figures

<sup>&</sup>lt;sup>b</sup> Selected model in bold; scaled residuals for selected model for doses 0, 125, and 250, were -8.73E-04, 5.93E-07, and 4.52E-07, respectively.

<sup>&</sup>lt;sup>c</sup> Gamma has the lowest AIC when considering five significant figures (72.988) vs. the Probit model that had an AIC of 72.998.

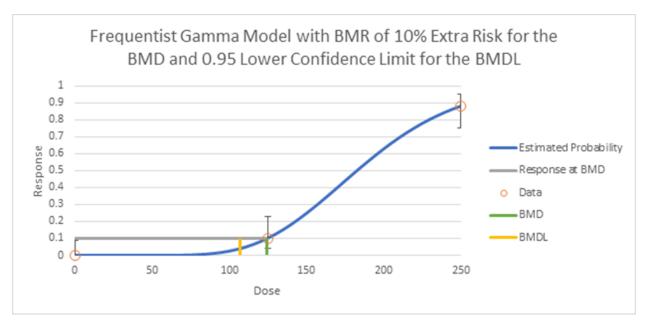


Figure 1-32. Plot of Mean Response by Dose with Fitted Curve for the Selected Model (Gamma) for Renal Tubule Karyomegaly (Nuclear Enlargement) in Female Mice Exposed to TCEP Via Oral Gavage in mg/kg/day; BMR 10 Percent Extra Risk

Model Results					
Benchma					
BMD	124.5420284	_			
BMDL	107.145575				
BMDU	139.3151775				
AIC	72.98782296				
P-value	0.999303735				
D.O.F.	1				
Chi²	7.615E-07				
Model Pa	rameters				
# of Parameters		3			
Variable	Estimate				
g	Bounded	7			
a	12.91059878	$\dashv$			
b	0.068833285	-			
	0.000033203	_			
		_			
Goodne	ss of Fit				
Dose	Estimated	Expected	Observed	Size	Scaled
Dose	Probability	Expected	Observed	Size	Residual
0	1.523E-08	7.61499E-07	0	50	-0.000873
125	0.102040791	4.999998743	5	49	5.932E-07
250	0.879999979	43.99999896	44	50	4.524E-07
	'				•
Analysis of	Deviance				
Model	Log Likelihood	# <u>of</u> Parameters	Deviance	Test d.f.	P Value
Full Model	-34.49391072	3	-	-	NA
Fitted Model	-34.49391148	2	1.523E-06	1	0.9990153
Reduced Model	-94.37178638	1	119.755751	2	<0.0001
			•	•	•

Figure 1-33. Details Regarding the Selected Model (Gamma) for Renal Tubule Karyomegaly (Nuclear Enlargement) in Female Mice Following Oral Exposure to TCEP in a Two-Year Chronic Toxicity Study

### 1.1.3.5 Relative Kidney Weight in Female Rats

Relative kidney weights increased in female mice exposed to TCEP for 16 weeks (NTP, 1991b). For BMD modeling, the administered doses were first duration adjusted to estimate an equivalent oral dose for animals exposed for seven days per week rather than five days per week. Then, continuous models were used to fit dose-response data.

A BMR of one SD were chosen according to EPA's *BMD Technical Guidance* (U.S. EPA, 2012). EPA did not identify a specific magnitude of change in relative kidney weight (*e.g.*, 10 percent) that would be considered biologically significant. The doses and response data used for the modeling are presented in Table 1-25.

Table 1-25. Female Rat Relative Kidney Weights and Associated Doses Selected for Dose-Response Modeling for TCEP From a 16-Week Study<sup>a</sup>

Dose (mg/kg-day)	Number of Animals	Mean	SD
0	10	3.69	0.13
16	8	3.83	0.17
31	10	4.03	0.13
63	10	4.1	0.22
125	8	4.18	0.17

<sup>&</sup>lt;sup>a</sup> The following data for the top dose of 250 mg/kg-day was not used: 5 animals, mean and SD of 4.51 and 0.13.

Table 1-26 summarizes the BMD modeling results for increased relative kidney weight in female rats in the 16-week study. For the full data set (using all dose groups), none of the available models provided adequate fit to the means (test 4 p-value < 0.1). Survival was decreased at the highest dose and EPA considered that the models could be run using the control and four lower doses. Although data are not available on the cause of all the deaths, two females died after receiving double doses for three days and several of the overdosed animals; the cause of deaths of three other female rats was not stated. Without the highest dose, the constant variance model provided adequate fit to the variance data (test 2 p-values > 0.05) and with the model applied, the Exponential 4 and 5 models provided adequate fit to the means (test 4 p-value > 0.1). The BMDLs for the fit models were sufficiently close (differed by < 3-fold); therefore, EPA selected the model with the lowest AIC (Exponential 4).

Table 1-26. Summary of BMD Modeling Results for Increased Relative Kidney Weights in Female Rats Following Oral Exposure to TCEP in a 16-Week Study (Highest Dose Group Dropped; Constant Variance Assumed)<sup>a b</sup>

	Goodness of Fit (Means)		BMD 1SD	BMDL 1SD	Basis for Model	
Model	Test 4 p-value	AIC	(mg/kg-day)	(mg/kg-day)	Selection	
Exponential 2	0.00430	-20.2	51.9	39.5		
Exponential 3	0.00430	-20.2	51.9	39.5		
<b>Exponential 4</b>	0.496	-30.0	12.5	7.41	The Exponential 4	
Exponential 5	0.297	-28.3	16.9	7.64	model is recommended	
Hill	0.448	-28.8	16.5	7.02	because it provided adequate fit to the	
Polynomial Degree 4	0.00569	-20.8	49.0	36.9	means (test 4 p-value > 0.1) and resulted in	
Polynomial Degree 3	0.00569	-20.8	49.0	36.9	the lowest AIC.	
Polynomial Degree 2	0.00569	-20.8	49.0	36.9		
Power	0.00569	-20.8	49.0	36.9		

	Goodness of Fit (Means) BMD 1SD		RMD 1SD	BMDL 1SD	Basis for Model
Model	Test 4 p-value	AIC	(mg/kg-day)	(mg/kg-day)	Selection Selection
Linear	0.00569	-20.8	49.0	36.8	

<sup>&</sup>lt;sup>a</sup> Three significant figures

 A plot of the Exponential 4 model with a BMR of 1 SD is shown in Figure 1-34. Additional modeling details, including model parameters, goodness of fit at each dose and log likelihood are shown in Figure 1-35.

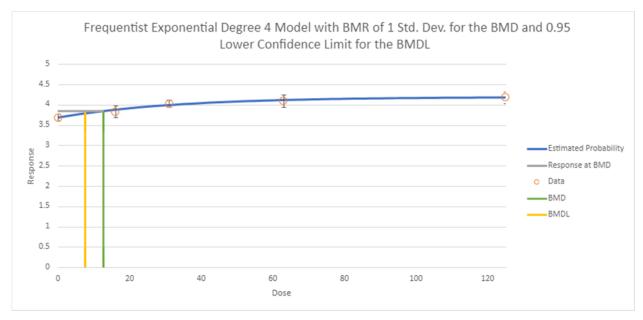


Figure 1-34. Plot of Response by Dose with Fitted Curve for the Selected Model (Exponential 4) for Relative Kidney Weight Increases in Female Rats Exposed to TCEP Via Oral Gavage (16-Week Study) and BMR of 1SD (Constant Variance Assumed)

<sup>&</sup>lt;sup>b</sup> Selected model in bold; scaled residuals for selected model for doses 0, 13, 31, 63, and 125 mg/kg-day were 0.176, -0.808, 0.790, -0.271, and 0.0317, respectively.

Model R	esults	<b>;</b>							
			1						
	enchma	rk Dose							
BMD		12.54183054							
BMDL		7.407226907							
BMDU		23.99667514							
AIC		-29.96772649							
Test 4 P-va	lue	0.495858212							
D.O.F.		2							
Mo	odel Par	rameters	1						
# <u>of</u> Param	eters	4							
Variab	le	Estimate	1						
а		3.68108785	1						
b		0.030132406	1						
С		1.13824347	1						
			,						
Goodnes	s of Fit								
Dose	Size	Estimated	Calc'd			Estimated	000000	Observed	
DOSC	SIZC	Median	Median	Mea		SD	SD	SD	Residual
0	10	3.68108785	3.69	3.6	9	0.16015224	0.13	0.13	0.175974389
									-
16	8	3.87575017	3.83	3.8	3	0.16015224	0.17	0.17	0.807987573
31	10	3.99001365	4.03	4.0	3	0.16015224	0.13	0.13	0.78954838
									-
63	10	4.113734445	4.1	4.1	l	0.16015224	0.22	0.22	0.271192763
125	8	4.178202794	4.18	4.1	8	0.16015224	0.17	0.17	0.031740207
Likeli	ihoods	of Interest			_				
	.		# 0						
Mode	51	Log Likelihood*	Parame	eters		AIC			
A1		19.6853285	6		_	.370657			
A2		21.67448971	10		_	.348979			
A3		19.6853285	6		_	.370657			
fitted	t	18.98386324	4		_	.967726			
R		0.455422464	2		3.0	8915507			
Te	ests of I	nterest							
		-							
				16					
				gt	-				
1		42.4381345	8		_	0.0001			
2		3.978322428	4		0.4	0894754			
3		3.978322428	4		0.4	0894754			
4		1.402930511	2		0.4	9585821			
2		3.978322428 3.978322428	8 4 4	<u>df</u>	0.4	0894754 0894754			

Figure 1-35. Details Regarding the Selected Model (Exponential 4) for Relative Kidney Weight Increases in Female Rats Following Oral Exposure to TCEP in a 16-Week Toxicity Study

## 1.1.4 Liver Toxicity

EPA modeled liver effects when a pairwise change from controls and/or dose-response trend was evident in the data (*e.g.*, a statistically significant change was identified).

When modeling liver weight changes, the best measures are changes relative to body weight (to account for any changes that are primarily related to body weight changes). However, EPA modeled both relative and absolute liver weight changes in male rats at 66 weeks in the two-year cancer bioassay and in female rats and mice from 16-week studies (NTP, 1991b) because body weights didn't change or

because the percent change in relative liver weight was 30 percent greater than changes in body weight in female rats at 350 mg/kg-day after 16 weeks.

All modeled results from the NTP studies are presented except the relative liver weight changes in male rats at 66 weeks because neither the constant nor the non-constant variance models provided adequate fit to the variance data. The female rat data could not be modeled without dropping doses and therefore, EPA is not presenting these data. EPA also modeled decreased absolute liver weight in male ICR mice in a 35-day study (Chen et al., 2015) as a comparison with liver weight changes from other studies, but these results are not shown because none of the models provided adequate fits to the data either assuming constant or non-constant variance.

## 1.1.4.1 Eosinophilic Foci in Male Mice

Male mice exhibited an increase in eosinophilic liver foci after two years of exposure to TCEP (NTP, 1991b). As inputs to BMD modeling and for consistency across endpoints, administered doses were first duration adjusted to estimate an equivalent oral dose for animals exposed for seven days per week rather than five days per week. Then, dichotomous models were used to fit dose-response data.

EPA presents the BMDL based on a BMR of 10 percent ER from the best fit model and based on the severity of the endpoint and considering EPA's *BMD Technical Guidance* (<u>U.S. EPA, 2012</u>). The doses and response data used for the modeling are presented in Table 1-27.

Table 1-27. Male Mouse Eosinophilic Foci in Livers and Associated Doses Selected for Dose-Response Modeling for TCEP in the Two-Year Bioassay

Dose (mg/kg-day)	Number of Animals	Incidence
0	50	0
125	50	3
250	50	8

Table 1-28 summarizes the BMD modeling results for eosinophilic foci in male mice. The Log-logistic, Multistage 2- and 1- degree, Logistic, and Probit models all provided adequate fits to the data (chi-square p-value > 0.1). BMDLs among the fit models were sufficiently close (< 3-fold difference). Therefore, EPA chose the model with the lowest AIC – the Multistage 1-degree model.

Table 1-28. BMD Modeling Results for Eosinophilic Liver Foci in Male Mice in the Two-Year Bioassay  $^a\,^b$ 

Model	Goodness of Fit (Means)		BMD 10%ER	BMDL 10%ER	Basis for Model	
	p-value	AIC	(mg/kg-day)	(mg/kg-day)	Selection	
Dichotomous Hill	N/A	72.7	169	0	Of the models	
Gamma	N/A	72.7	178	108	with adequate fits (Log-logistic,	
Log-Logistic	0.999	70.7	178	104	Multistage 2- and 1-degree,	
Multistage 2	0.999	70.7	180	108	Logistic, and	

Model	Goodness of Fit (Means)		BMD 10%ER	BMDL 10%ER	Basis for Model
	p-value	AIC	(mg/kg-day)	(mg/kg-day)	Selection
Multistage 1	0.878	68.9	168	106	Probit models), EPA chose the
Weibull	N/A	72.7	178	108	model with the
Logistic	0.339	72.0	208	172	lowest AIC.
Log-Probit	N/A	76.9	244	0	
Probit	0.398	71.7	202	163	

<sup>&</sup>lt;sup>a</sup> Three significant figures

 Plots of the Multistage 2-degree model with BMR 10 percent ER is shown in Figure 1-36. Additional modeling details, including model parameters, goodness of fit at each dose and log likelihood are shown in Figure 1-37.

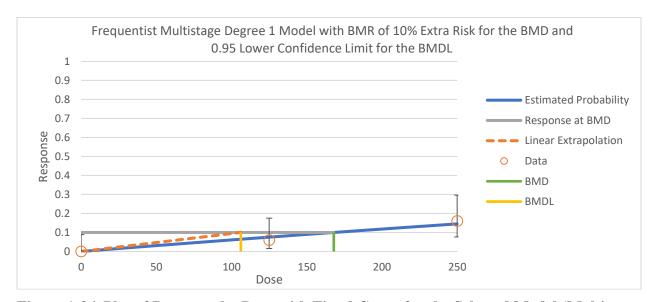


Figure 1-36. Plot of Response by Dose with Fitted Curve for the Selected Model (Multistage 1-Degree) for Eosinophilic Foci in Livers of Male Mice Exposed to TCEP Via Oral Gavage (Two-Year Bioassay) and BMR of 10 Percent

<sup>&</sup>lt;sup>b</sup> Selected model in bold; scaled residuals for selected model for doses 0, 125, and 250 mg/kg-day were –8.73E–04, –0.413, and 0.298 respectively.

M	ahr	l Resi	nits

Benchmark Dose					
BMD	167.9439247				
BMDL	106.14906				
BMDU	289.6913765				
AIC	68.93261182				
P-value	0.878392643				
D.O.F.	2				
Chi <sup>2</sup>	0.259323168				
Slope Factor	0.000942071				

Model Parameters				
# of Parameters	2			
Variable	Estimate			
g	Bounded			
b1	0.000627355			

Goodness of Fit					
Dose	Estimated Probability	Expected	Observed	Size	Scaled Residual
0	1.523E-08	7.61499E-07	0	50	-0.000873
125	0.075423454	3.771172705	3	50	-0.412992
250	0.145158198	7.257909888	8	50	0.2979257

Analysis of Deviance					
Model	Log Likelihood	# of Parameters	Deviance	Test d.f.	P Value
Full Model	-33.3318701	3	-	-	N/A
Fitted Model	-33.46630591	1	0.26887163	2	0.874209
Reduced Model	-39.32656941	1	11.9893986	2	0.0024919

Figure 1-37. Details Regarding the Selected Model (Multistage 1-Degree) for Eosinophilic Foci in Livers of Male Mice in the Two-Year Bioassay

# 1.1.4.2 Absolute Liver Weight in Male Rats

Absolute liver weights increased in male rats exposed to TCEP at 66 weeks (NTP, 1991b). As inputs to BMD modeling and for consistency across endpoints, administered doses were first duration adjusted to estimate an equivalent oral dose for animals exposed for seven days per week rather than five days per week. Then, continuous models were used to fit dose-response data.

BMRs of one SD and ten percent RD were chosen according to EPA's *BMD Technical Guidance* (U.S. EPA, 2012). EPA considers the BMR of ten percent RD to be adverse and used the BMDL associated with this BMR for consideration within the risk evaluation and when comparing with other PODs. The doses and response data used for the modeling are presented in Table 1-29.

10

Mean

14.9

16.2

17.9

SD

2.52

1.04

1.11

Dose (mg/kg-day)	Number of Animals
0	9

31

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Table 1-30 summarizes the BMD modeling results for increased absolute liver weight in male rats at 66 weeks in the NTP two-year chronic bioassay. Although the constant variance model did not provide adequate fit to the variance data, the non-constant variance model provided an adequate fit. With the non-constant variance model applied, the Exponential 2, Exponential 3, 2-degree Polynomial, Power and Linear models provided adequate fit to the means (test 4 p-value > 0.1). The BMDLs for the fit models with adequate fit were sufficiently close (< 3-fold difference). The Power and 2-degree Polynomial models converged on the Linear model; these had the lowest AICs and the Linear model was selected as the most parsimonious choice.

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Table 1-30. Summary of BMD Modeling Results for Increased Absolute Liver Weights in Male Rats Following Oral Exposure to TCEP at 66 Weeks (Non-constant Variance)<sup>a b</sup>

Model	Goodness of Fit		BMD 1SD	BMDL 1SD	BMD 10%RD	BMDL 10%RD	Basis for Model
lylodei	p-value	AIC	(mg/kg- day)	(mg/kg- day)	(mg/kg- day)	(mg/kg- day)	Selection
Exponential 2	0.587	109	45.0	30.3	32.2	23.0	Among the non-
Exponential 3	0.591	109	44.3	30.3	31.9	23.0	constant variance models with
Exponential 4	N/A	111	39.1	19.5	24.9	9.83	adequate fit (test 4
Exponential 5	N/A	111	39.0	19.5	24.8	9.83	p-value > 0.1), the Linear model is
Hill	65535	113	31.9	18.8	30.3	28.2	recommended because it is the
Polynomial 2	0.694	109	42.7	28.3	29.8	20.4	most parsimonious of the three
Power	0.694	109	42.7	28.3	29.8	20.4	converged models with the lowest
Linear	0.694	109	42.7	28.3	29.8	20.4	AICs.

<sup>&</sup>lt;sup>a</sup> Three significant figures

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Plots of the Linear model with BMRs of one SD and ten percent RD are shown in Figure 1-38 and Figure 1-39, respectively. Additional modeling details, including model parameters, goodness of fit at each dose and log likelihood are shown in Figure 1-40.

<sup>&</sup>lt;sup>b</sup> Selected model in bold; scaled residuals for selected model for doses 0, 31, and 63 mg/kg-day were 0.200, 0.216, and 0.0636, respectively.

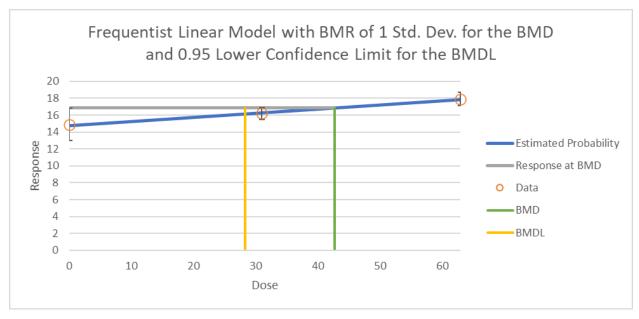


Figure 1-38. Plot of Response by Dose with Fitted Curve for the Selected Model (Linear) for Absolute Liver Weight Increases in Male Rats Exposed to TCEP Via Oral Gavage (at 66 Weeks) and BMR of 1SD (Non-constant Variance)

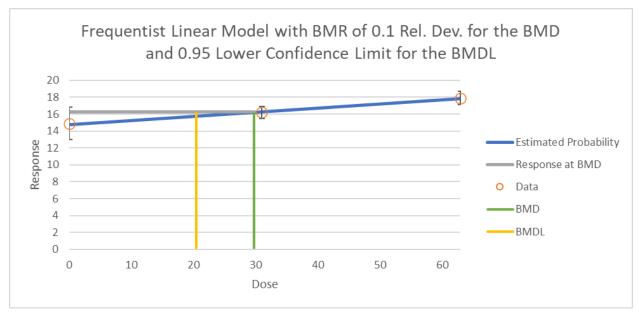


Figure 1-39. Plot of Response by Dose with Fitted Curve for the Selected Model (Linear) for Absolute Liver Weight Increases in Male Rats Exposed to TCEP Via Oral Gavage (at 66 Weeks) and BMR of 10 Percent (Non-constant Variance)

	Results	nork Doco							
BMD 42.69774628									
BMDL		28.32874416							
BMDU		76.64400531							
AIC		109.1787602							
	P-value	0.694137339							
D.O.F.		1							
D.O.I .		1							
	Model P	arameters							
# of Pa	rameters		4						
Va	riable	Estimate							
	g	14.75901536	;						
b	eta1	0.049556637	•						
	rho	-8.45314728							
		•							
	ness of								
	Fit	Estimated	Calc'd	Oh	served	Estimated	Calc'd	Observed	
Dose	Size		Median		Mean	SD	SD	SD	Scaled Residual
0	9	14.75901536	14.9		14.9	2.11595582	2.52	2.52	0.199887879
31		14.73701330	17./		17./				
31	10	16.29527111	16.2		16.2		1.04	1.04	
63	10	16.29527111 17.88108349	16.2 17.9			1.39234274 0.94032713		1.04	-0.216378973
63	10	17.88108349			16.2 17.9	1.39234274	1.04		
63	10	17.88108349 Is of Interest	17.9		17.9	1.39234274 0.94032713	1.04	1.04	-0.216378973
63 N	10  Likelihood  Iodel	17.88108349  Is of Interest  Log Likelihood	17.9 d* # of	Para	17.9	1.39234274 0.94032713	1.04	1.04	-0.216378973
63 M	10 Likelihood Iodel A1	17.88108349 Is of Interest	17.9 d* # of	Para	17.9	1.39234274 0.94032713	1.04	1.04	-0.216378973
63 M	10 Likelihood Iodel A1 A2	17.88108349  Is of Interest  Log Likelihood	17.9 d* # of	Para 4	17.9	1.39234274 0.94032713	1.04	1.04	-0.216378973
63 M	10 Likelihood Iodel A1	17.88108349  Is of Interest  Log Likelihood  -54.25967213	17.9 d* # of 3	Para 4 6 5	17.9	1.39234274 0.94032713 AIC 116.519344	1.04	1.04	-0.216378973
63 N	10 Likelihood Iodel A1 A2	17.88108349 Is of Interest Log Likelihood -54.25967213 -49.31972604	17.9  d* # of  3  4	Para 4 6 5 4	17.9	AIC 116.519344 110.639452 111.024117 109.17876	1.04	1.04	-0.216378973
63 N	10  Likelihood  Iodel  A1  A2  A3	17.88108349 Is of Interest Log Likelihoo -54.25967213 -49.31972604 -50.5120586	17.9  d* # of  3  4  7	Para 4 6 5	17.9	AIC 110.639452 111.024117	1.04	1.04	-0.216378973
63 N	Likelihood lodel A1 A2 A3 itted R	17.88108349  Is of Interest Log Likelihood -54.25967213 -49.31972604 -50.51205866 -50.58938012 -61.09564682	17.9  d* # of  3  4  7	Para 4 6 5 4	17.9	AIC 116.519344 110.639452 111.024117 109.17876	1.04	1.04	-0.216378973
63 M	Likelihood lodel A1 A2 A3 litted R	17.88108349  Is of Interest Log Likelihood -54.25967213 -49.31972604 -50.51205866 -50.58938012 -61.09564682	17.9  d* # of 3  4  7  2  2	Para 4 6 5 4 2	17.9	AIC 116.519344 110.639452 111.024117 109.17876 126.191294	1.04	1.04	-0.216378973
63 M	IO  Likelihood  Iodel  A1  A2  A3  itted  R	17.88108349  Is of Interest Log Likelihood -54.25967213 -49.31972604 -50.51205867 -50.58938012 -61.09564682  Ests of Interest -2*Log(Likeli	17.9  14* # of 3  4 7  2 2	Para 4 6 5 4 2	17.9 meters	AIC 116.519344 110.639452 111.024117 109.17876 126.191294  p-value	1.04	1.04	-0.216378973
63 M	In the state of th	17.88108349  Is of Interest Log Likelihood -54.25967213 -49.31972604 -50.51205867 -50.58938012 -61.09564682  Ests of Interest -2*Log(Likelide) 23.5518	17.9  d* # of  3 4  7 2  2 bhood Rat  34155	Para 4 6 5 4 2	Test df	AIC 116.519344 110.639452 111.024117 109.17876 126.191294  p-value <0.0001	1.04	1.04	-0.216378973
63 M	IO  Likelihood  Iodel  A1  A2  A3  itted  R	17.88108349  Is of Interest Log Likelihood -54.25967213 -49.31972604 -50.51205867 -50.58938012 -61.09564682  Ests of Interest -2*Log(Likeli	17.9  17.9  18.4  19.5	Para 4 6 5 4 2	17.9 meters	AIC 116.519344 110.639452 111.024117 109.17876 126.191294  p-value	1.04	1.04	-0.216378973

Figure 1-40. Details Regarding the Selected Model (Linear) for Absolute Liver Weight Increases in Male Rats at 66 Weeks

### 1.1.4.3 Absolute Liver Weight in Female Mice

Absolute liver weights increased in female mice exposed to TCEP for 16 weeks (NTP, 1991b). For BMD modeling, the administered doses were first duration adjusted to estimate an equivalent oral dose for animals exposed for seven days per week rather than five days per week. Then, continuous models were used to fit dose-response data.

BMRs of one SD and ten percent RD were chosen according to EPA's *BMD Technical Guidance* (U.S. EPA, 2012). EPA considers the BMR of ten percent RD to be adverse and used the BMDL associated with this BMR for consideration within the risk evaluation and when comparing with other PODs. The doses and response data used for the modeling are presented in Table 1-31.

Table 1-31. Female Mouse Absolute Liver Weights and Associated Doses Selected for
 Dose-Response Modeling for TCEP From a 16-Week Study

Dose (mg/kg-day)	Number of Animals	Mean	SD
0	10	1.07	0.09
31	10	1.11	0.13
63	10	1.16	0.09
125	9	1.22	0.12
250	9	1.29	0.12
500	10	1.21	0.06

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Table 1-32 summarizes the BMD modeling results for increased absolute liver weight in female mice in the 16-week study. The constant variance model provided adequate fit to the variance data. With the constant variance model applied, the Exponential 4 and 5 models and the Hill model provided adequate fit to the means. The BMDLs for these models were sufficiently close (< 3-fold difference). Therefore, EPA selected the Exponential 4 model because it has the lowest AIC.

Table 1-32. Summary of BMD Modeling Results for Increased Absolute Liver Weights in Female Mice Following Oral Exposure to TCEP in a 16-Week Study (Constant Variance)<sup>a b c</sup>

Model	Goodness of Fit		BMD 1SD	BMDL 1SD	BMD 10%RD	BMDL 10%RD	Basis for Model
lylodel	p-value	AIC	(mg/kg- day)	(mg/kg- day)	(mg/kg- day)	(mg/kg- day)	Selection
Exponential 2	0.00233	-81.9	447	294	446	291	
Exponential 3	0.00233	-81.9	447	294	447	292	The Exponential 4 and 5 models and the
<b>Exponential 4</b>	0.268	-92.5	57.8	27.7	61.5	28.0	Hill model provided
Exponential 5	0.211	-91.4	71.2	31.2	75.7	32.3	adequate fit to the means (test 4 p-
Hill	0.174	-91.0	68.9	31.3	73.0	36.3	values > 0.1). The BMDLs for the fit
Polynomial 5	0.00269	-82.2	428	275	428	273	models were sufficiently close (<
Polynomial 4	0.00269	-82.2	428	275	428	273	3-fold difference). Therefore, EPA
Polynomial 3	0.00269	-82.2	428	275	428	273	chose the Exponential 4 model,
Polynomial 2	0.00269	-82.2	428	275	428	273	which has the lowest AIC.
Power	0.00269	-82.2	428	276	428	273	AIC.
Linear	0.00269	-82.2	428	275	428	273	

<sup>&</sup>lt;sup>a</sup> Three significant figures

<sup>&</sup>lt;sup>b</sup> Based on test 2 p-values > 0.05 for all models, EPA determined that the constant variance model assumption may be suitable for dose-response modeling.

Model	Goodness of Fit		BMD 1SD	<b>DIVID</b>		BMDL 10%RD	Basis for Model
	p-value	AIC	(mg/kg- day)	(mg/kg- day)	(mg/kg- day)	(mg/kg- day)	Selection

<sup>&</sup>lt;sup>c</sup> Selected model in bold; scaled residuals for selected model for doses 0, 31, 63, 125, 250, and 500 mg/kg-day were 0.311, 0.442, 0.236, 0.201, 1.43, and 1.18, respectively.

Plots of the Exponential 4 model with BMRs of one SD and 10 percent RD are shown in Figure 1-41 and Figure 1-42, respectively. Additional modeling details, including model parameters, goodness of fit at each dose and log likelihood are shown in Figure 1-43.

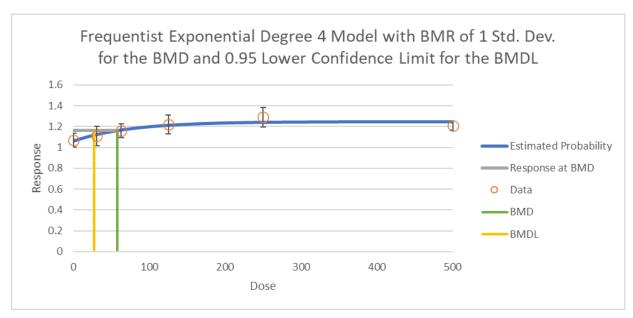


Figure 1-41. Plot of Response by Dose with Fitted Curve for the Selected Model (Exponential 4) for Absolute Liver Weight Increases in Female Mice Exposed to TCEP Via Oral Gavage (16-Week Study) and BMR of 1SD (Constant Variance)

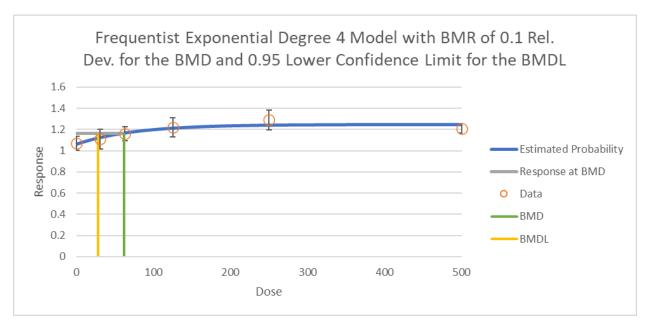


Figure 1-42. Plot of Response by Dose with Fitted Curve for the Selected Model (Exponential 4) for Absolute Liver Weight Increases in Female Mice Exposed to TCEP Via Oral Gavage (16-Week Study) and BMR of 10 Percent Relative Deviation (Constant Variance)

Model Results	
Benchma	ark Dose
BMD	57.75678158
BMDL	27.71499956
BMDU	163.5531039
AIC	-92.52408587
Test 4 P-value	0.267822421
D.O.F.	3

Model Pa	rameters
# of Parameters	4
Variable	Estimate
a	1.059984909
b	0.013471904
С	1.177474954

Goodne	ss of Fit	]						
Dose	Size	Estimated Median	Galc'd Median	Observed Mean	Estimated SD	SD SD	Observe d SD	Scaled Residual
0	10	1.059984909	1.07	1.07	0.1017205 6	0.09	0.09	0.31134806 9
31	10	1.124208116	1.11	1.11	0.1017205 6	0.13	0.13	- 0.44170038 9
63	10	1.167597871	1.16	1.16	0.1017205 6	0.09	0.09	- 0.23620179 5
125	9	1.21318441	1.22	1.22	0.1017205 6	0.12	0.12	0.20100922 1
250	9	1.24162317	1.29	1.29	0.1017205 6	0.12	0.12	1.42675675 5
500	10	1.2478823	1.21	1.21	0.1017205 6	0.06	0.06	1.17768084

Likelihoods	of Interest		
Model	Log Likelihood*	# <u>of</u> Parameters	AIC
A1	52.23292131	7	-90.465843
A2	55.72879398	12	-87.457588
A3	52.23292131	7	-90.465843
fitted	50.26204294	4	-92.524086
R	39.66391724	2	-75.327834

Tests of	Interest		
Test	2* <u>Log(</u> Likelihood Ratio)	Test df	p-value
1	32.12975349	10	0.00038098
2	6.991745356	5	0.22125494
3	6.991745356	5	0.22125494
4	3.941756742	3	0.26782242

Figure 1-43. Details Regarding the Selected Model (Exponential 4) for Absolute Liver Weight Increases for Female Mice Exposed in a 16-Week Study

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#### 1.1.4.4 Relative Liver Weight in Female Mice

Relative liver weights increased in female mice exposed to TCEP for 16 weeks (NTP, 1991b). For BMD modeling, the administered doses were first duration adjusted to estimate an equivalent oral dose for animals exposed for seven days per week rather than five days per week. Then, continuous models were used to fit dose-response data.

BMRs of one SD and 10 percent RD were chosen according to EPA's BMD Technical Guidance (U.S. EPA, 2012). EPA considers the BMR of ten percent RD to be adverse and used the BMDL associated with this BMR for consideration within the risk evaluation and when comparing with other PODs. The doses and response data used for the modeling are presented in Table 1-33.

Table 1-33. Female Mouse Relative Liver Weights and Associated Doses Selected for Dose-Response Modeling for TCEP From a 16-Week Study

Dose (mg/kg-day)	Number of Animals	Mean	SD
0	10	41.5	3.64
31	10	41.7	5
63	10	42.8	4.02
125	9	45.9	3.69
250	9	48.6	4.05
500	10	47.4	3.29

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1017 1018 Table 1-34 summarizes the BMD modeling results for increased relative liver weight in female mice in the 16-week study. The Exponential 4 and 5 models and the Hill model provided adequate fit to the means (test 4 p values > 0.1) using the constant variance model. The BMDLs for these models differed by less than 3-fold, and therefore, EPA chose the Exponential 5 model because it has the lowest AIC.

Table 1-34. Summary of BMD Modeling Results for Increased Relative Liver Weights in Female Mice Following Oral Exposure to TCEP in a 16-Week Study (Constant Variance)<sup>a b c</sup>

Model	Goodness of Fit		BMD 1SD	BMDL 1SD	BMD 10%RD	BMDL 10%RD	Basis for Model	
	p-value	AIC	(mg/kg- day)	(mg/kg- day)	(mg/kg- day)	(mg/kg- day)	Selection	
Exponential 2	0.0349	335	334	240	344	247	The Exponential 4	
Exponential 3	0.0349	335	334	240	344	247	and 5 models and the Hill model	
Exponential 4	0.409	330	89.7	44.4	96.7	46.2	provided adequate fit to the means (test	
Exponential 5	0.783	329	112	61.0	119	64.4	4 p values > 0.1). The BMDLs for the	
Hill	0.706	329	109	61.3	116	69.2	fit models were sufficiently close	
Polynomial 5	0.0419	335	317	223	327	229	(differed by < 3-	
Polynomial 4	0.0419	335	317	223	327	229	fold); therefore, EPA chose the	

Model	Goodness of Fit		BMD 1SD	BMDL 1SD	BMD 10%RD	BMDL 10%RD	Basis for Model	
Wiodei	p-value	AIC	(mg/kg- day)	(mg/kg- day)	(mg/kg- day)	(mg/kg- day)	Selection	
Polynomial 3	0.0419	335	317	223	327	229	Exponential 5 model, which had	
Polynomial Degree 2	0.0419	335	317	223	327	229	the lowest AIC.	
Power	0.0419	335	317	223	327	229		
Linear	0.0419	335	317	223	327	229		

<sup>&</sup>lt;sup>a</sup> Three significant figures

 Plots of the Exponential 5 model with BMRs of one SD and ten percent RD are shown in Figure 1-44 and Figure 1-45, respectively. Additional modeling details, including model parameters, goodness of fit at each dose and log likelihood are shown in Figure 1-46.

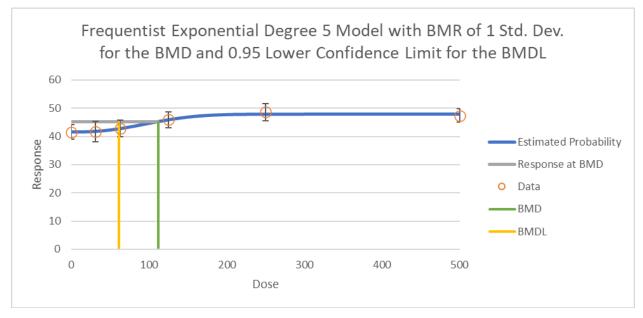


Figure 1-44. Plot of Response by Dose with Fitted Curve for the Selected Model (Exponential 5) for Relative Liver Weight Increases in Female Mice Exposed to TCEP Via Oral Gavage (16-Week Study) and BMR of 1SD (Constant Variance)

<sup>&</sup>lt;sup>b</sup> Based on test 2 p-values > 0.05 for all models, EPA determined that the constant variance model assumption may be suitable for dose-response modeling.

<sup>&</sup>lt;sup>c</sup> Selected model in bold; scaled residuals for selected model for doses 0, 31, 63, 125, 250, and 500 mg/kg-day were 0.00298, 0.0292, 0.0424, 0.0332, 0.512, and 0.470, respectively.

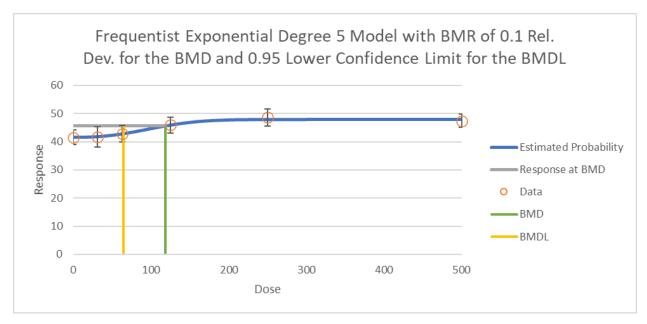


Figure 1-45. Plot of Response by Dose with Fitted Curve for the Selected Model (Exponential 5) for Relative Liver Weight Increases in Female Mice Exposed to TCEP Via Oral Gavage (16-Week Study) and BMR of 10 Percent Relative Deviation (Constant Variance)

Model Results	
Benchm	nark Dose
BMD	111.7706299
BMDL	60.96094969
BMDU	189.8088568
AIC	329.2304296
Test 4 P-value	0.783232337
D.O.F.	2

Model Pa	rameters
# of Parameters	5
Variable	Estimate
а	41.49642547
b	0.008505426
c	1.155854346
d	2.461378592

Goodne	ss of Fit	]						
Dose	Size	Estimated Median	€aJ¢'d Median	Observed Mean	Estimated SD	Gale'd SD	Observe d SD	Scaled Residual
0	10	41.49642547	41.5	41.5	3.7925859 2	3.64	3.64	0.00298046
31	10	41.7349856	41.7	41.7	3.7925859 2	5	5	- 0.02917117 3
63	10	42.74919598	42.8	42.8	3.7925859 2	4.02	4.02	0.04236065
125	9	45.94195855	45.9	45.9	3.7925859 2	3.69	3.69	- 0.03318992 6
250	9	47.95311831	48.6	48.6	3.7925859 2	4.05	4.05	0.51169442 5
500	10	47.96382372	47.4	47.4	3.7925859 2	3.29	3.29	- 0.47011911 6

Likelihood	s of Interest		
Model	Log Likelihood*	# <u>of</u> Parameters	AIC
A1	-159.3708889	7	332.741778
A2	-158.306178	12	340.612356
A3	-159.3708889	7	332.741778
fitted	-159.6152148	5	329.23043
R	-171.9698877	2	347.939775

Tests of	Interest		
Test	- 2* <u>Log(</u> Likelihood Ratio)	Test df	p-value
1	27.32741956	10	0.00231098
2	2.129421912	5	0.83096277
3	2.129421912	5	0.83096277
4	0.488651801	2	0.78323234

Figure 1-46. Details Regarding the Selected Model (Exponential 5) for Relative Liver Weight Increases for Female Mice Exposed in a 16-Week Study

## 1.2 Cancer

EPA modeled endpoints for kidney tumors, the only tumors that had robust evidence if one or more doses resulting in pairwise differences from controls and/or if a dose-response trend was evident in the two-year cancer bioassay (NTP, 1991b). Evidence for tumors at other target organs was slight. The BMD/BMDLs chosen for tumor incidence were based on animals still alive at the time the first incidence of cancer was observed. Also, preference was given to presenting BMD models that included both adenomas and carcinomas because benign tumors (adenomas) are expected to lead to malignant tumors (carcinomas).<sup>2</sup>

EPA did not present BMD modeling after combining tumors from multiple target organs, because the combinations would include tumors for which evidence was slight.

## 1.2.1 Renal Tubule Adenomas and Carcinomas (Combined) in Male Rats

Male rats exhibited increased incidences of renal tubule carcinomas and adenomas in the two-year NTP bioassay (NTP, 1991b). For BMD modeling, the administered doses were first duration adjusted to estimate an equivalent oral dose for animals exposed for seven rather than five days per week. Then, two Multistage models were used to fit dose-response data.

 EPA chose a BMR of 10 percent ER to model the tumor data according to EPA's *Benchmark Dose Technical Guidance* (U.S. EPA, 2012). The doses and response data used for the modeling using both kidney adenomas and carcinomas are presented in Table 1-35. The numbers of animals were adjusted for mortality. Specifically, the modeling included only the animals still alive when the first tumor was observed (day 575).

Table 1-35. Male Rat Renal Tubule Adenomas or Carcinomas (Combined) and Associated Doses Selected for Dose-Response Modeling for TCEP from a Two-Year Chronic Bioassay

Dose (mg/kg-day)	Number of Animals	Incidence <sup>a</sup>
0	40	2
31	44	5
63	44	25
<sup>a</sup> Increased incidence of carcin	noma was identified – 1 control	and 1 high-dose rat

 Table 1-36 summarizes the BMD modeling results for combined renal tubule adenomas and carcinomas in male rats. EPA selected the 2-degree Multistage model because it was the only model that provided an adequate fit (chi-square p-value > 0.1) to the data.

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<sup>&</sup>lt;sup>2</sup> As a comparison, EPA also conducted BMD modeling of tumor incidence from an 18-month dietary study using ddY mice (<u>Takada et al., 1989</u>) (not shown). Tumors included: Renal cell adenomas and carcinomas in males; hepatocellular adenomas and carcinomas in males; leukemia in females; and forestomach papillomas and squamous cell carcinomas in females. <u>Takada et al. (1989)</u> is in a foreign language and was not critical to using quantitatively in the risk evaluation; furthermore, EPA did not evaluate it for data quality. One or more of the multistage models fit each of these tumor type/sex combinations but ddY mice were less sensitive than the species used by <u>NTP (1991b)</u> based on the resulting cancer slope factors (CSFs).

Table 1-36. Summary of BMD Modeling Results for the Combined Incidence of Renal Tubule Adenomas and Carcinomas in Male Rats Following Oral Exposure to TCEP in a Two-Year Chronic Bioassay $^{a\,b}$ 

**BMD BMDL CSF Goodness of Fit** 10% ER 10%ER (per **Basis for Model Selection** Model (mg/kg-(mg/kgmg/kg-day) p-value AIC day) day) EPA chose the 2-degree Multistage 2 0.144 114 24.6 17.2 0.0058 Multistage model because it was the only model that provided an adequate fit (chisquare p-value > 0.1) to the 0.00439 8.83  $ND^{c}$ Multistage 1 120 12.1 data

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1082 1083 EPA also modeled adenomas alone and identified a CSF of  $6.0 \times 10^{-3}$  per mg/kg-day but considered the slope factor based on both adenomas and carcinomas to be the most appropriate for the risk evaluation. A plot of the Multistage 2 model with a BMR of 10 percent ER is shown in Figure 1-47. Additional modeling details, including model parameters, goodness of fit at each dose, and log likelihood are shown in Figure 1-48.

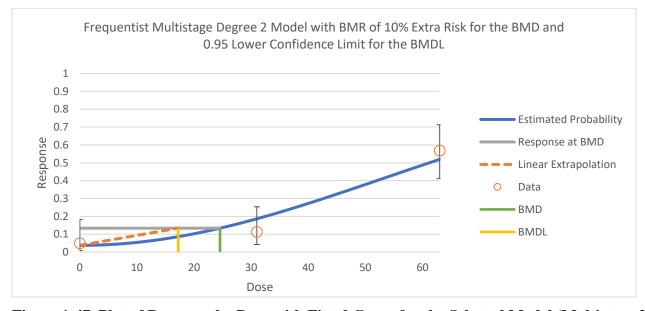


Figure 1-47. Plot of Response by Dose with Fitted Curve for the Selected Model (Multistage 2) for the Combined Incidence of Renal Tubule Adenomas and Carcinomas in Male Rats

<sup>&</sup>lt;sup>a</sup> Three significant figures

<sup>&</sup>lt;sup>b</sup> Selected model in bold; scaled residuals for selected model for doses 0, 31, and 63 mg/kg-day were 0.408, -0.124, and 0.652, respectively.

<sup>&</sup>lt;sup>c</sup> ND = Not determined

Model Results	
Benchr	mark Dose
BMD	24.55384094
BMDL	17.23177476
BMDU	29.64335493
AIC	113.527872
P-value	0.144414026
D.O.F.	1
Chi <sup>2</sup>	2.130283692
Slope Factor	0.005803233
Model F	Parameters
# of Parameters	3
Variable	Estimate

Woderrara	illeters
# of Parameters	3
Variable	Estimate
g	0.037717037
b1	Bounded
bz	0.000174759

Goodness	of Fit				
Dose	Estimated Probability	Expected	Observed	Size	Scaled Residual
0	0.037717037	1.508681474	2	40	0.4077678
31	0.186484185	8.205304142	5	44	-1.24062
63	0.519084789	22.83973072	25	44	0.6518207

Analysis of	Deviance				
Model	Log Likelihood	# of Parameters	Deviance	Test d.f.	P Value
Full Model	-53.60696754	3	-		NA
Fitted Model	-54.76393601	2	2.31393695	1	0.1282189
Reduced Model	-71.97889851	1	36.7438619	2	<0.0001

Figure 1-48. Details Regarding the Selected Model (Multistage 2) for the Combined Incidence of Renal Tubule Adenomas and Carcinomas in Male Rats

#### **1.2.2** Renal Tubule Adenomas in Female Rats

Female rats exhibited increased incidences of renal tubule adenomas in the two-year NTP bioassay (NTP, 1991b). For BMD modeling, the administered doses were first duration adjusted to estimate an equivalent oral dose for animals exposed for seven rather than five days per week. Then, two Multistage models were used to fit dose-response data.

EPA chose a BMR of 10 percent ER to model the tumor data according to EPA's *Benchmark Dose Technical Guidance* (U.S. EPA, 2012). The doses and response data used for the modeling are presented in Table 1-37. The numbers of animals were adjusted for mortality. Specifically, the modeling included only the animals still alive when the first tumor was observed (day 729).

Table 1-37. Female Rat Renal Tubule Adenomas and Associated Doses Selected for Dose-Response Modeling for TCEP from Two-Year Chronic Bioassay

Dose (mg/kg-day)	Number of Animals	Incidence <sup>a</sup>
0	32	0
31	33	2
63	17	5
<sup>a</sup> Female rats had no renal tubulo	e carcinomas.	

 Table 1-38 summarizes the BMD modeling results for renal tubule adenomas in female rats. Both Multistage models provided an adequate fit to the data (chi-square p-value > 0.1), and the BMDLs for the models were sufficiently close (< 3-fold difference). Therefore, EPA selected the Multistage 2 model, which had the lowest AIC.

Table 1-38. Summary of BMD Modeling Results for the Incidence of Renal Tubule Adenomas in Female Rats Following Oral Exposure to TCEP in a Two-Year Chronic Bioassay $^ab$ 

Model	Goodnes	ss of Fit	BMD 10%ER (mg/kg-	BMDL 10% ER (mg/kg-	CSF (per mg/kg-	Basis for Model Selection
	p-value	AIC	day)	day)	day)	
Multistage 2	0.938	37.8	36.3	19.3	0.0052	Both models provided an adequate fit (chi-square p-value > 0.1), and the BMDLs
Multistage 1	0.213	41.3	28.6	16.2	$\mathrm{ND}^c$	were sufficiently close (< 3- fold difference). Thus, EPA chose the Multistage 2 model, which had the lowest AIC.

 $<sup>^</sup>a$  Three significant figures

A plot of the Multistage 2 model with a BMR of 10 percent ER is shown in Figure 1-49. Additional modeling details, including model parameters, goodness of fit at each dose and log likelihood are shown in Figure 1-50.

<sup>&</sup>lt;sup>b</sup> Selected model in bold; scaled residuals for selected model for doses 0, 31, and 63 mg/kg-day were -0.000698, -0.290, and 0.211, respectively.

<sup>&</sup>lt;sup>c</sup> ND = Not determined

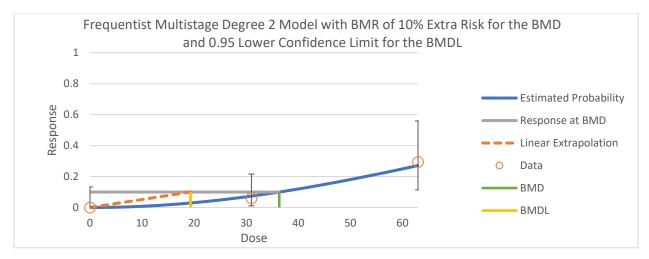


Figure 1-49. Plot of Response by Dose with Fitted Curve for the Selected Model (Multistage 2) for the Incidence of Renal Tubule Adenomas in Female Rats

		1			
Benchma	rk Dose				
BMD	36.34603715				
BMDL	19.30952154				
BMDU	51.52675798				
AIC	37.81956123				
P-value	0.937802873				
D.O.F.	2				
Chi <sup>2</sup>	0.128431017				
Slope Factor	0.005178792				
Model Par	ameters				
# <u>of</u> Parameters	3				
Variable	Estimate				
g	Bounded				
b1	Bounded				
b2	7.97561E-05				
		1			
Goodnes	s of Fit				
Dose	Estimated Probability	Expected	Observed	Size	Scaled Residual
0	1.523E-08	4.87359E-07	0	32	-0.000698
31	0.073781956	2.434804551	2	33	-0.289538
63	0.27134279	4.612827437	5	17	0.2111832
Analysis of	Deviance				
Model	Log Likelihood	# of Parameters	Deviance	Test d.f.	P Value
Full Model	-17.84340932	3	-	-	NA
Fitted Model	-17.90978062	1	0.1327426	2	0.9357833
Reduced Model	-23.91799872	1	12.1491788	2	0.0023006

Figure 1-50. Details Regarding the Selected Model (Multistage 2) for the Incidence of Renal Tubule Adenomas in Female Rats

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